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Author(s): Kivisaari, Kati; Boratyński, Zbyszek; Lavrinienko, Anton; Kesäniemi, Jenni; Lehmann, Philipp; Mappes, Tapio

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The effect of chronic low-dose environmental radiation on organ mass of bank voles in the Chernobyl exclusion zone

Kati Kivisaari^a , Zbyszek Boratyński^b , Anton Lavrinienko^a , Jenni Kesäniemi^a , Philipp Lehmann^c , and Tapio Mappes^a 

^aDepartment of Biological and Environmental Science, University of Jyväskylä, Jyväskylä, Finland; ^bCIBIO/InBIO, Research Center in Biodiversity and Genetic Resources, University of Porto, Vairão, Portugal; ^cDepartment of Zoology, Stockholm University, Stockholm, Sweden

ABSTRACT

Purpose: Animals are exposed to environmental ionizing radiation (IR) externally through proximity to contaminated soil and internally through ingestion and inhalation of radionuclides. Internal organs can respond to radioactive contamination through physiological stress. Chronic stress can compromise the size of physiologically active organs, but studies on wild mammal populations are scarce. The effects of environmental IR contamination on organ masses were studied by using a wild rodent inhabiting the Chernobyl exclusion zone (CEZ).

Material and methods: The masses of brain, heart, kidney, spleen, liver and lung were assessed from bank voles (*Myodes glareolus*) captured from areas across radioactive contamination gradient within the CEZ. Relative organ masses were used to correct for the body mass of an individual.

Results: Results showed a significant negative correlation between IR level in the environment and relative brain and kidney mass. A significant positive correlation between IR and relative heart mass was also found. Principal component analysis (PCA) also suggested positive relationship between IR and relative spleen mass; however, this relationship was not significant when spleen was analyzed separately. There was no apparent relationship between IR and relative liver or lung mass.

Conclusions: Results suggest that in the wild populations even low but chronic doses of IR can lead to changes in relative organ mass. The novelty of these result is showing that exposure to low doses can affect the organ masses in similar fashion as previously shown on high, acute, radiation doses. These data support the hypothesis that wildlife might be more sensitive to IR than animals used in laboratory studies. However, more research is needed to rule out the other indirect effects such as radiosensitivity of the food sources or possible combined stress effects from e.g. infections.

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Background

Radioactive disasters and nuclear fallouts are a major global concern. Long-lived radioactive compounds, such as cesium-137 (Cs^{137}) and strontium-90 (Sr^{90}) from Fukushima and Chernobyl, have spread far in landscape scale and due to their long half-life (approximately 30 years) the effects persist for decades (Askbrant et al. 1996; IAEA 1996; Baba 2013). In Chernobyl area, many species gain most of their radiation burden externally and accumulation of Cs^{137} and Sr^{90} is species-specific (Beresford et al. 2020). Due to its water solubility, Cs^{137} has high mobility and enters the body through the diet and drinking water (Murakami et al. 2015). However, both Cs^{137} and Sr^{90} can also enter the body as radioactive particles from soil and air (Chesser et al. 2001). The rate of accumulation from external sources is usually expressed as dose rate of Sieverts per hour (Sv/h) and

annual radiation dose from natural sources on average is 2.4 mSv/year (approximately 0.27 $\mu\text{Sv/h}$) (IAEA 2010). The ambient radiation levels measured from the Chernobyl area can exceed 300 $\mu\text{Sv/h}$, thus accumulated doses could potentially cause cellular damage (Waselenko et al. 2004). The absorbed dose is expressed in Grays (Gy) which in case of gamma and beta active radioisotopes (both having radiation weighting factor of 1) would be converted 1:1 from Sieverts (IAEA 2010). Given the risks of future radiation-related accidents (Wheatley et al. 2016), studies investigating the potential environmental and health consequences of chronic exposure to chronic low doses (less than 2 Gy/year) of radionuclides especially in natural conditions are warranted.

The effects of ionizing radiation (IR) can be both direct (e.g. cellular) and indirect (e.g. oxidative stress) and both can cause damage in organs (Navarrete et al. 2011; Azzam et al.

CONTACT Kati Kivisaari  kati.kivisaari@jyu.fi; Tapio Mappes  tapio.mappes@jyu.fi 

 Supplemental data for this article can be accessed [here](#).

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2012). It is well established that chronic high doses of IR can cause also indirect radiation effects through radiation-induced oxidative stress (reaction to increased levels of reactive oxygen species), direct alteration of cell structures, and DNA damage (Azzam et al. 2012). Organs such as the heart, small intestine, liver and kidneys are energetically costly and make up a high proportion of basal metabolic rate (Konarzewski and Diamond 1995). In previous research, it has been shown that Cs^{137} indeed accumulates more in organs like the gastrointestinal tract and liver when fed orally to mice (Nelson et al. 1961) and in gastrointestinal tract, heart, kidney and liver when injected to dogs (Leggett et al. 2003), implying that metabolically active organs gain higher doses of Cs^{137} . In addition, Cs^{137} has been shown to increase concentration in higher stages of the food chain and its biological half-life often increases with the body mass (Pendleton et al. 1965; Brooks et al. 2016). Doses used in many previous radiation studies testing organs have been high and acute (e.g. total doses of >10 Gy in liver cancer treatment) (Tao and Yang 2012; Braunstein and Nakamura 2013; Emami 2013). Furthermore, most studies on radiation and tissue damage are also conducted in laboratory conditions, which are not necessarily ecologically relevant. Wildlife studies, focusing on IR induced organ damage and the effects of oxidative stress effects on cells, are mostly conducted on invertebrates, plants and birds, but rarely on mammals, the most suitable model for human radiosensitivity tests (Garnier-Laplace et al. 2013). For example, in wild birds, a negative IR effect on relative brain size has been shown (Møller et al. 2011), but only few similar wildlife studies have been conducted using small mammals. There is some evidence about molecular-level damage in Chernobyl area, since testes and liver have been shown to have shorter telomers, whereas brain and liver have higher telomerase expression in Chernobyl bank voles (Kesäniemi, Lavrinienko, Tukalenko, Boratyński, et al. 2019). Recent reviews suggested that wild populations are more radiosensitive than those used in laboratory studies, probably due to multiple stressors in their natural environment (Garnier-Laplace et al. 2013; Mothersill et al. 2019). Extrapolating the effects of IR, observed in the laboratory setting, on wild populations is not straightforward, due to the complexity of natural systems (Mothersill et al. 2019), yet highlighting the importance of wildlife studies in addition to controlled experiments.

Previous studies indicate that relative organ masses (accounted for variation in individuals size) are reliable indicators of individual exposure to toxicity (Sellers et al. 2007; Piao et al. 2013). It has been proposed that radiation-induced organ damage can be visualized as decrease in organ mass due to inhibition of growth or cell death (Lumniczky et al. 2017). Here, we tested if major visceral organs (brain, heart, spleen, kidneys and lungs) differ in their responses to IR in the environment in a wild muroid rodent, the bank vole (*Myodes [= Clethrionomys] glareolus*, Schreber 1780). Experimental animals were captured from the Chernobyl exclusion zone (CEZ) in Ukraine (Figure 1) across a range of radioactive contamination. Energetically costly organs such as the brain and liver were expected to

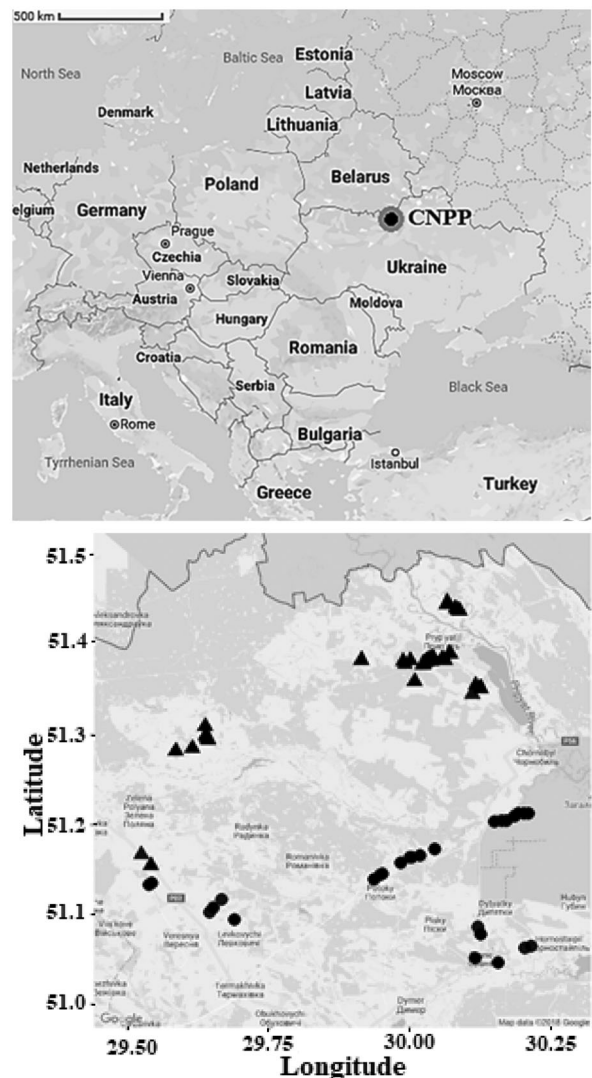


Figure 1. Map of trapping locations surrounding Chernobyl nuclear power plant (CNPP) located in northern Ukraine. Map on the right shows our trapping areas from all trappings. Radiation level in contaminated sites (triangles) varied between 1.46 and 98.74 $\mu\text{Gy/h}$ and in uncontaminated sites (dots) 0.01–0.53 $\mu\text{Gy/h}$.

show strong negative responses to radioactive contamination (Aiello and Wheeler 1995; Navarrete et al. 2011). The liver and spleen are also involved in detoxification and the functioning immune system, thus these tissues are known to suffer from high level of exposures to contaminants (Miller et al. 2005; Toesca et al. 2018). Kidneys are responsible for filtration of blood and are, therefore, exposed to all water-soluble products, including Cs^{137} , and thus were also expected to suffer consequences of elevated internal radiation exposure (Nelson et al. 1961). Due to the limited proliferative capacity (Bhattacharya and Asaithamby 2016) and high radiotolerance (Darby et al. 2010), strong changes in heart mass in response to contamination were not expected.

Material and methods

Study animals

Bank voles are widespread throughout Europe, found in both coniferous and deciduous forests (Ledevine et al. 2010).

Bank voles are the most abundant rodent species in the CEZ and can be found even in the most contaminated locations (Chesser et al. 2000), however, the abundances decrease with increasing radiation in the capture locations (Mappes et al. 2019). Bank voles have very variable diet from roots, plant stems and leaves to seeds, lichens and insects (Canova 1993). Experimental animals were captured from a total of 41 trapping sites within and just outside the CEZ in June 2011, 2016 and 2017 (Figure 1). Since the only fraction of adults survives two winters (Prévot-Julliard et al. 1999), most of the captured adult individuals were likely born in previous autumn, being similar age during the experiment. Sites were selected to cover environments with variable radiation levels and distances from Chernobyl Nuclear Power Plant (CNPP; Figure 1). All areas had optimal conditions for bank voles and possible sub-populations were considered by sampling from both sides of geographical barriers (i.e. rivers). As previous research has found no effect of environmental radiation on ground level plant coverage (0–50 cm), the most important vegetation variable for bank voles (Mappes et al. 2019), no large differences in food sources among our trapping locations were not expected. We used Ugglan Special2 live traps (Grahnb, Sweden) baited with potato and sunflower seeds and distance between neighboring trapping sites was at least 1 km. In 2011, each trapping site had 20 traps, whereas 2016–2017 trappings, each site had 16 traps. Traps were checked daily and captured animals were transported into the laboratory located in the CEZ.

Radiation was measured next to each trap from ground level (no more than 1 cm above the ground) using a hand-held radiation detector (Gamma-Scout w/ALERT Radiation detector/Geiger Counter, Gamma-Scout GmbH & Co. KG, Germany, with built-in shielding to block alpha radiation). This resulted in 20 readings per site in 2011 and 16 readings per site in 2016–2017, which were averaged to get estimated environmental radiation level ($\mu\text{Gy/h}$) for each trapping site (Supplementary Table 1). Radiation data were right skewed, so the radiation measurements were \log_{10} -transformed to achieve normality. A recent study shows that approximately 70% of radiation dose of Chernobyl bank voles comes from external sources of Cs^{137} and Sr^{90} (Beresford et al. 2020). However, IR also accumulates in plant tissue, which is eaten by animals and increase their exposure (Boratyński et al. 2016). The actual dose of each individual can thus have large variance, even within sites depending on ingested food sources, nevertheless individual doses are still strongly correlated with environmental radiation measurements from trapping sites (Chesser et al. 2000; Garnier-Laplace et al. 2013). Our measurements from the CEZ show a similar correlation (Supplementary Figure 1). Some estimates of internal Cs^{137} burden for our study species have been published previously, stating that in animals captured from the highly contaminated areas of the CEZ Cs^{137} burden was up to 11,678,718.6 Bq/kg (Lavrinenko et al. 2018). Trapping sites were considered to be uncontaminated if averaged ambient radiation level was less than 1 $\mu\text{Gy/h}$. Radiation level in contaminated sites where bank voles were successfully captured

(within 30 km exclusion zone (EZ), including the most contaminated locations within 10 km distance from CNPP) varied between 1.46 and 198.74 $\mu\text{Gy/h}$, and in uncontaminated sites (within 30 km EZ and outside the EZ) between 0.01 and 0.58 $\mu\text{Gy/h}$ (Figure 1; Supplementary Table 2).

In the following experiments, we used adult bank voles ($N=221$) captured around the CNPP (Figure 1). Juvenile individuals were excluded based on visual traits (grey juvenile fur) and body mass (<14 g). After transportation to the laboratory, captured individuals were sacrificed through cervical dislocation, sexed and their body mass, body length and head width were recorded. All organs were inspected for possible visual signs of infections (e.g. lesions), but none of the organ samples showed signs of infection. Different stages of pregnancy in females were accounted for by subtracting the embryo mass from whole body mass. Internal organs (brain, heart, liver and spleen) were dissected, cleaned of fat and connective tissue, blotted dry and weighted with an analytic balance (Mettler Toledo, XS105 DualRange, precision: 0.01 g). From a subset of captured individuals (individuals from 2016) also kidneys ($N=162$) and lungs ($N=109$) were collected.

Ethical statement

All experiments complied with the legal requirements and adhered closely to international guidelines for the use of animals in research. All necessary permissions were obtained from the Animal Experimentation Committee for these experiments (permission no. ESAVI/3834/04.10.03/2011 and ESAVI/7256/04.10.07/2014).

Statistics

In the initial check of the raw data, simple correlations were found between radiation (\log_{10} -transformed to normalize right-skewed data) and organ masses (Supplementary Table 3). There was also significant negative correlation between radiation and brain, liver and kidney mass (Supplementary Table 3). However, heart, spleen and lungs did not follow the same pattern. To infer the patterns of variation in organ masses, a principal component analysis (PCA) was conducted with Promax rotation and Kaiser Normalization since all organ masses are highly correlated. PCA is a multivariate technique that reduces the number of correlated traits to a smaller number of uncorrelated variables (Jolliffe 2002). The first PCA revealed only one principal component, where over 50% of the variation in organ masses was explained likely by body mass (Supplementary Table 4). Thus, the overall effect of radiation on body mass was tested with a generalized linear mixed model (GLMM) where body mass (\log_{10} -transformed) was included as dependent variable, radiation (\log_{10} -transformed) and sex were included as explanatory variables, and trapping site was included as a random factor (Supplementary Table 5). A significant negative correlation was found between radiation and body mass (Supplementary Table 5), however, this was expected since larger animals usually have larger organs. Before statistical

analyses, normality and homogeneity of variance in traits were confirmed. Statistical significance was established at the .05 level.

Since body mass was highly correlated with organ masses (Supplementary Figure 2), this variation in body mass was taken into account by calculating standardized residual organ masses (standardized residuals from the linear regression between organ mass and body mass). The second PCA was then conducted with these corrected variables (from here on referred to as residual organ mass). The number of considered principal components was selected so that cumulative variances exceeded 70% and eigenvalues >0.8 . This second PCA analysis divided the residual organ masses along three principal components (PC1, PC2 and PC3; Table 1), which subsequently were used in the GLMMs to assess the putative effects of radiation. The Akaike information criterion (AIC) scores were used to compare and rank four different models including (i) only radiation level (as covariate; \log_{10} -transformed), (ii) only sex, (iii) both sex and radiation level and (iv) both sex and radiation level and their interaction. Trapping site (instead of the trapping year) was included as random factor in all models since trapping areas varied intentionally between years, to cover larger areas, and only part of the sites (roughly one third) was overlapping among years. Models with body mass included as covariate were also tested to confirm that body mass did not correlate with residual organ masses. Since the effect was non-significant in all cases ($t < 0.841$, $p > .401$) and residual organ masses already controlled for body mass, we excluded body mass from the final models. The age effect was also tested by adding head width as the proxy of the age to the model (Kallio et al. 2014). This had no significant effect on any of the organs, and since adult individuals were intentionally selected for the experiment, the control for age was left out to simplify the model. Two models were assumed equally informative when difference in AIC scores between them is lower than 2.00 (Burnham and Anderson 2002). The best model included both sex and radiation level and controlled for trapping site as random factor:

$$\text{Principal component} \sim \text{Radiation} (\log_{10}) + \text{Sex} + (1 | \text{trapping site}) \quad (1)$$

Table 1. Principal component analysis (PCA) with residual organ masses (brain, heart, liver and spleen mass ($N=221$), corrected for body mass).

Organ	PC1	PC2	PC3
Eigenvalue	1.389	1.065	0.818
Percent of variance	34.89	26.74	20.53
Cumulative % of var.	34.89	61.64	82.18
Brain	0.000	0.962	0.000
Heart	0.598	-0.197	0.150
Liver	-0.002	0.000	0.982
Spleen	0.655	0.178	-0.136

Number of considered principal components was selected so that cumulative variances $\geq 70\%$ and eigenvalues were >0.8 . Loadings with value >0.50 are highlighted in bold. Extraction method in PCA was applied with Promax with Kaiser Normalization since nearly all organ masses are highly correlated.

A subset of samples included also masses of kidneys ($N=162$) and lungs ($N=109$). According to Mundfrom et al. (2005), the sample size of 109 would not be sufficient to pass excellent level criterion (K values at least 0.98) with six variables to >4 factors. Therefore, to maximize the sample size in the PCA, radiation effects on each organ mass were tested separately (to include subset organs) with the following model:

$$\text{Residual organ mass} \sim \text{Radiation} (\log_{10}) + \text{Sex} + (1 | \text{trapping site}) \quad (2)$$

Since older animals potentially have been exposed to radioactive environment for longer, test was also repeated using only individuals with body mass larger than 20 g, which gave qualitatively similar results (Supplementary Table 6). Statistical analyses were conducted using SPSS version 24.0.0.1 (Chicago, IL) and RStudio (Integrated Development for R. RStudio, PBC, Boston, MA) version 1.0.143 (R Core Team 2016) with following packages: lme4, nlme. Packages ggmap (Kahle and Wickham 2013) and ggplot2 (Wickham 2009) were used for map graphics.

Results

Principal component analysis

Body and internal organ mass were measured from 221 bank voles (subset of $N=162$ included kidneys and $N=109$ included lungs) and ambient radiation level was recorded from each trapping location (descriptive statistics in Supplementary Table 2). A PCA with original organ masses (Supplementary Table 4) produced a single principal component (PC) accounting over 50% of variation, likely representing the overall individual size. Thus, when testing radiation effects on bank vole body mass, a significant negative relationship was found between body mass and radiation (GLMM: coefficient = -0.743 , $t = -3.231$, $df = 217$, $p = .001$; Supplementary Table 5). There was no significant difference in body mass between two sexes (Supplementary Table 5). Also, the number of males and females did not statistically differ between contaminated and uncontaminated areas (Binary logistic regression: $\chi^2 = 0.010$, $df = 1$, $p = .920$). Since contaminated areas tended to host smaller individuals (Supplementary Figure 4), and as all organ masses were positively correlated with body mass (Pearson correlation: $R^2 > 0.087$, $p < .001$), residual values of organ masses corrected for animal body mass (linear regression: organ mass \sim body mass) were calculated. The subsequent PCA using these residual organ masses (Table 1) grouped organs ($N=221$ including corrected brain, heart, liver and spleen masses) into three components accounting for over 80% of variation. The first component (PC1) had loadings in heart and spleen, PC2 in brain and PC3 in liver (Table 1). These three PCs were then used in subsequent analyses to test for radiation effects on organ masses.

Testing for the effects of radiation

Analysis of the three principal components in relation to ambient radiation level, sex and trapping location revealed significant effect of radiation in two out of three models (Table 2). A significant positive relationship was found between PC1 (loadings in heart and spleen masses; Table 2) and radiation. On the contrary, PC2 (loadings in brain) showed a strong negative relationship with radiation and a significant sex difference (as overall brain mass in females was significantly smaller). Finally, PC3 (loadings in liver) showed neither significant relationship with radiation nor sex differences. Subset of samples included additional data for kidney ($N = 162$) and lung ($N = 109$) masses. These tests were run for all residual organ masses separately (see statistics and Table 3). The analyses showed very similar results as previous analyses using PCA scores (Table 2). Radiation had a significant negative effect on brain mass, but also on kidney mass (not included in the PCA). Heart mass was positively correlated with radiation (Table 3). The masses of lung, spleen and liver did not correlate with radiation when analyzed separately. There were also few sex differences: the masses of brain, kidney and spleen were significantly smaller in female voles (Table 3). In addition, a model using older individuals with body mass larger than 20 g was tested. The results were very similar to those presented in Table 2 (Supplementary Table 6), thus we had no valid reason to exclude smaller animals from the dataset.

Discussion

Radioactive cesium (Cs^{137}) easily spreads throughout the body due to its water-soluble properties and can be

Table 2. Radiation effects on principal components ($N = 221$). Radiation is \log_{10} -transformed ambient radiation level from trapping site. Statistically significant values are highlighted in bold.

Dependent	Fixed	Coef.	Std. err	t	p Value
PC1 (Heart and spleen)	Radiation	0.251	0.071	3.535	<.001
	Sex (female)	-0.194	0.132	-1.475	.142
PC2 (brain)	Radiation	-0.280	0.063	-4.456	<.001
	Sex (female)	-0.651	0.115	-5.642	<.001
PC3 (liver)	Radiation	-0.086	0.072	-1.197	.223
	Sex (female)	0.101	0.135	0.455	.233

Table 3. Results from mixed model analyses with organ masses corrected by body mass.

Dependent	Fixed	Coef.	Std. err	t	p Value
Brain	Radiation	-0.272	0.066	-4.123	<.001*
	Sex (female)	-0.564	0.118	-4.773	<.001*
Heart	Radiation	0.295	0.071	4.132	<.001*
	Sex (female)	0.162	0.130	1.247	.214
Liver	Radiation	-0.113	0.073	-1.556	.121
	Sex (female)	0.013	0.135	0.099	.921
Spleen	Radiation	0.128	0.072	1.778	.077
	Sex (female)	-0.444	0.131	-3.383	.001*
Kidneys ($N = 162$)	Radiation	-0.248	0.076	-3.283	.001*
	Sex (female)	-0.346	0.146	-2.368	.019
Lungs ($N = 109$)	Radiation	0.062	0.092	0.679	.499
	Sex (female)	-0.097	0.196	-0.493	.623

For brain, heart, liver and spleen $N = 221$, for kidneys $N = 162$ and for lungs $N = 109$ (see methods for statistical analyses). Statistically significant results are highlighted in bold.

*Effects significant after Bonferroni correction, for alpha = 0.008.

dangerous as both external gamma source as well as internal beta emitter (Nelson et al. 1961). However, the accumulation of Cs^{137} in different parts of the body in mice varies significantly and after a single dose, excretion reduces the concentration in most organs to fractions of the imbibed dose in mere days (Nelson et al. 1961). Since the bank voles inhabiting the CEZ constantly live in a contaminated environment, their chronic doses are expected to covary with contamination levels within the CEZ (Chesser et al. 2000). The variation in organ masses was assessed across areas with varying levels of ambient radiation, from uncontaminated areas to areas where yearly doses can raise up to 2 Gy/year (habitats where ambient radiation level exceeds 200 $\mu\text{Gy/h}$).

Previous research has shown that in areas with higher radiation, bank voles have shorter telomeres and higher increase in telomerase expression in brain and liver (Kesäniemi, Lavrinienko, Tukalenko, Boratyński, et al. 2019), thus these organs were expected to also show stronger negative morphological responses to radioactive contamination (Aiello and Wheeler 1995; Navarrete et al. 2011). Telomere length has been linked to cell aging (Shay and Wright 2000), and this could be one potential mechanisms behind radiation effects seen in the bank voles in the CEZ. Consistent with our expectations, a significant decrease in residual brain mass with increasing radiation level was found in bank voles exposed to chronic low dose radiation. In laboratory animals, damage in brain cells and smaller brain has been linked to increased prevalence of neurological aberrations, such as cognitive dysfunctions, and in humans to mental retardation and lower intelligence (Hladik and Tapio 2016; Verreet et al. 2016; Lumniczky et al. 2017). However, doses used in laboratory settings are often very high (total doses >50 Gy) and poorly comparable to those measured in wild animal populations in the CEZ (Lumniczky et al. 2017). In laboratory mice, prenatal doses of 0.66 Gy have been shown to affect brain mass of the offspring, but only a dose of 1 Gy started to affect body mass (Verreet et al. 2016). In addition, brain has poor regenerative capacity and radiation injury from young age can persist into adulthood (Lumniczky et al. 2017). In Chernobyl, decreased brain size has been observed in birds, indicating that chronic radiation could affect organs at lower doses than observed in laboratory studies (Møller et al. 2011). A recent study shows that proportional head size of the fetuses of Japanese monkeys has decreased significantly following the Fukushima disaster, also implying a decrease in brain size (Hayama et al. 2017). Radiosensitivity of the fetus is also well established in laboratory studies (Brown 1964; Devi and Hande 1990).

Kidneys are responsible for filtration of blood and are, therefore, exposed to water-soluble products, including Cs^{137} , and thus were expected to be affected by elevated internal radioactive exposure (Nelson et al. 1961). Inhaled Cs^{137} is transported quickly through lungs and accumulates mostly in kidneys, skeletal muscle and intestines (Nelson et al. 1961). There was a significant negative relationship between environmental radiation level and relative kidney mass. High dose IR (5–20 Gy) has been reported to inhibit renal growth in weanling mice (Donaldson et al. 1978). It

has also been shown that reduction in kidney mass can lead to hypertension or even renal failure (Williams 1986). To our knowledge, there are no previous studies assessing the effect of very low dose radiation on kidney mass. Since Cs^{137} acts in similar fashion with potassium when entering to animal body and is transported with plasma, a large portion of it is filtered through kidneys and liver (Leggett et al. 2003). Kidneys accumulate 20–25% of Cs^{137} (Stather 1970) and Leggett et al. (2003) estimated kidneys to have highest Cs^{137} concentration in the body when radioactive solution is injected. Being a tissue with slow turnover rate (Williams 1986), kidney could also be affected by cell bystander effects (where unirradiated cells exhibit irradiated effects as a result of signals received from irradiated cells) and suffer from cumulative radiation injury (Mothersill et al. 2017). It can be speculated that organs with higher turnover and regeneration rates do not reflect this effect; yet testing this hypothesis remains for experiments dedicated for this purpose.

Heart has a limited proliferative capacity (Bhattacharya and Asaithamby 2016) and has been shown to have high radiotolerance (Darby et al. 2010). Even though single doses as high as 15 Gy have been shown to cause radiation-induced inflammation followed by thickening of heart muscle (Darby et al. 2010), heart mass was not expected to increase in response to lower-level radioactive contamination measured from the CEZ. Regardless of our expectations, a positive correlation between environmental radiation and relative heart mass was found. The heart has been considered to be radioresistant in humans and has been thought to be unaffected by doses below 30 Gy; but recent studies are showing marked effects also with doses below 20 Gy (Darby et al. 2010). The effects of the chronic low-dose (<2 Gy) radiation exposure on heart tissue are still poorly understood (Taunk et al. 2015). The heart muscle has low antioxidant defense and radiation-induced oxidative stress could cause cellular damage (Tapio 2016). Radiotherapy in humans has been reported to cause both enlargement of heart and thickening of heart muscle (Yusuf et al. 2011) and also hypertension (Souza et al. 2015). Thickening of heart muscle also increases the pumping pressure and can thus lead to an overall enlargement of the heart (Darby et al. 2010). Both thickening and enlargement can result in heart problems and even heart failure (Yusuf et al. 2011). On the contrary, Sreetharan et al. (2019) showed that acute doses of 1 Gy caused reduction in the heart rate of male rat offspring when mother was irradiated during pregnancy. However, again the doses given in radiation therapy and laboratory studies are much larger than those estimated for bank voles from Chernobyl (Chesser et al. 2000). In addition, radiation-induced effects in heart often appear months or years after the exposure (Williams et al. 2010; Emami 2013), thus it is possible that chronic exposure could cause similar effects. Furthermore, the heart tissue of newborn mice has been shown to have developmental alterations even in low single doses of 0.02 Gy and these persisted to adulthood (Bakshi et al. 2013). Therefore, it is plausible that cellular damage gained in early life or in utero has persisted to adult

individuals also in the CEZ and contributes to the higher heart mass observed in animals from contaminated habitats.

The spleen is involved in the immune system responses, thus is known to have negative response to high level of contaminants (Miller et al. 2005; Toesca et al. 2018). There was a positive relationship between relative spleen mass and radiation. Similar increase in spleen mass, few days after irradiation with large doses (>8 Gy), has been previously observed in mice (Congdon and Urso 1957). This increase was associated with proliferation of cells and appeared in response to increased infection rate and pressure to immune system (Makinodan et al. 1962). Radiation effects in spleen appear usually after considerable delay due to slow turnover of cells (Williams 1986). IR can directly damage cells (including blood cells) and the spleen is responsible for both storing white blood cells and recycling of old blood cells (Osipov et al. 2013; Ochiai et al. 2014). Increase in spleen size could thus result from a vicious cycle of radiation-induced white cell damage and pressure to the immune system. This hypothesis is also supported by a recent study, where the expression of genes involved in inflammatory responses was observed to be upregulated in Chernobyl bank voles (Kesäniemi, Jernfors, et al. 2019). In [Supplementary Figure 2](#), larger (and therefore older) individuals have high variation in their spleen mass. Frequently occurring infections could possibly increase spleen mass (Ponlet et al. 2011). Bank voles are often infected with different types of hantaviruses, but these have mainly asymptomatic effects on their host (Kallio et al. 2010). Previous study done with Chernobyl bank voles did not find any difference in viral prevalence or viral load with bank voles captured from within and outside the CEZ (Kesäniemi, Lavrinienko, Tukalenko, Mappes, et al. 2019). With previous data, there are also no differences in the body conditions of the bank voles captured from within and outside the CEZ indicating that there are no cases of malnutrition (Kesäniemi, Jernfors, et al. 2019). It is, however, possible, that some combined stress effects from these would cause stronger effects in some organs.

Lungs of animals inhabiting the CEZ are potentially exposed to IR as any inhaled radioactive particles first reach the lungs (Stone et al. 2003), but no statistically significant effect of radiation on the lungs was found. There was no significant effect of radiation on liver either, which could indicate that both organs have a higher tolerance against chronic radiation doses. The liver is considered to have relatively high radio resistance and in humans doses up to 30 Gy have been considered safe in radiotherapy (Emami 2013; Benson et al. 2016). Doses of this magnitude could be gained only in few locations in the CEZ and only within decades (e.g. the Red Forest, ambient radiation level >200 $\mu\text{Gy/h}$), thus it is practically impossible for the bank vole to get very high doses even during their whole lifetime. The radio-resistance of lungs has been described to be lower than that of the liver, but no clear threshold has been established so far (Emami 2013). The longer the half-life of the radioactive material, the more likely that a significant proportion of radionuclides will move from lungs to

surrounding tissue and blood before decaying (Kendall and Smith 2002). In the case of Cs¹³⁷, hardly any activity is measured from mice lungs and instead, activity quickly moves to blood and accumulates in skeletal muscle (Nelson et al. 1961). Bank voles in the CEZ have been living in environment with chronic radiation exposure over 50 generations (Baker et al. 2017) and, therefore, it is possible that some form of radioresistance adaptation has evolved.

In addition to radiation effect, both PC2 (loadings in brain) and residual brain mass revealed significant sex difference, with significantly smaller brain mass in females. Since there was no interaction between radiation and sex effect, this effect is likely not related to contamination in the CEZ. Smaller relative brain size in females has also been shown in humans (Ankney 1992). Sex difference in brain size between males and females could be explained by higher levels of oxidative stress that are usually observed in females (Alonso-Alvarez et al. 2004). Both IR and free radicals increase the oxidative damage in cells (Alonso-Alvarez et al. 2004; Speakman 2008; Azzam et al. 2012) and in mammals, during lactation metabolism increases and generates even more free radicals (Speakman 2008). When using residual organ masses in the model, similar sex difference also in kidneys and spleen (females had significantly smaller kidney and spleen mass) was found, but again no interaction between radiation and sex effects. The sex differences were also absent with relative heart, liver and lung mass. Reproductive costs in females could explain the smaller organ size, since bank vole females can have up to ten pups in one litter (Koskela et al. 1998). A link between radiation and reproductive output was shown in a previous study where ambient radiation level was negatively correlated with female litter size (Lehmann et al. 2016; Mappes et al. 2019).

Trapped bank voles were significantly smaller in areas with increased ambient radiation level, which indicates that individuals are either younger or grow slower in contaminated areas. It is possible that voles facing high contamination do not survive as long as voles inhabiting lower contamination areas, where radiation level is close to natural background level. Alternatively, growth rate could be lower in areas with higher radiation exposure, which could explain the smaller body mass of bank voles in the more contaminated locations. Decreased growth rate and smaller body size have indeed been observed in butterflies (Hiyama et al. 2013) and delayed growth rate has been seen in monkey fetuses (Hayama et al. 2017) in the contaminated areas of Fukushima. The bank vole litter sizes have been shown to decrease with increasing radiation in the CEZ (Mappes et al. 2019), indicating some reproductional costs in contaminated sites. Our data suggest that low environmental radiation can induce changes in organ masses, particularly in organs with slow cell turnover rate. Unfortunately, in our study, it is not possible to separate the damage that occurred in young individuals and persisted into adulthood from the effects that are due to combined dose of the whole lifetime of the individual. Chronic laboratory studies normally use adult animals, that are more radioresistant than young individuals (Spalding and Trujillo 1962), and exposure does not last for whole lifetime. By using a wild mammal that is closely related

to common laboratory models (mice and rats), we show that tolerance against chronic IR in the environment varies among organs, and some organs (e.g. brain, heart, spleen and kidneys) might be more radiosensitive than previously expected (Emami 2013). Here, for the first time with wild mammal, we show that lifetime chronic exposure to chronic low-level radiation can have an effect on organ masses similarly to effects seen with high radiation doses. Thus, our work improves the understanding of how environmental contamination can affect wild organism and highlights the importance of wildlife studies to assess the effects of radiation in natural conditions. However, more research is still needed to assess the biological importance of these effects.

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Notes on contributors

Kati Kivisaari, PhD, is a project researcher in the Department of Biological and Environmental Science of the University of Jyväskylä, Finland.

Zbyszek Boratyński, PhD, is a postdoctoral researcher in CIBIO-InBIO, Research Center in Biodiversity and Genetic Resources, University of Porto, Portugal.

Anton Lavrinienko, MSc, is a PhD student in the Department of Biological and Environmental Science of the University of Jyväskylä, Finland.

Jenni Kesäniemi, PhD, is a postdoctoral researcher the Department of Biological and Environmental Science of the University of Jyväskylä, Finland.


Philipp Lehmann, PhD, is a postdoctoral researcher in Department of Zoology, Stockholm University, Sweden.

Tapio Mappes, PhD, Senior Lecturer in the Department of Biological and Environmental Science of the University of Jyväskylä, Finland.

ORCID

Kati Kivisaari  <http://orcid.org/0000-0001-8892-2947>

Zbyszek Boratyński  <http://orcid.org/0000-0003-4668-4922>

Anton Lavrinienko  <http://orcid.org/0000-0002-9524-8054>

Jenni Kesäniemi  <http://orcid.org/0000-0001-8328-558X>
 Philipp Lehmann  <http://orcid.org/0000-0001-8344-6830>
 Tapio Mappes  <http://orcid.org/0000-0002-5936-7355>

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