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Biogeochemical cycling and ecological thresholds in a High Arctic lake (Svalbard)

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Abstract

Lakes are a dominant feature of the Arctic landscape and a focal point of regional and global biogeochemical cycling. We collected a sediment core from a High Arctic Lake in southwestern Svalbard for multiproxy paleolimnological analysis. The aim was to find linkages between the terrestrial and aquatic environments in the context of climate change to understand centennial-long Arctic biogeochemical cycling and environmental dynamics. Two significant thresholds in elemental cycling were found based on sediment physical and biogeochemical proxies that were associated with the end of the cold Little Ice Age and the recent warming. We found major shifts in diatom, chironomid and cladoceran communities and their functionality that coincided with increased summer temperatures since the 1950s. We also discovered paleoecological evidence that point toward expanded bird (Little Auk) colonies in the catchment alongside climate warming. Apparently, climate-driven increase in glacier melt water delivery as well as a prolonged snow- and ice-free period have increased the transport of mineral matter from the catchment, causing significant water turbidity and disappearance of several planktonic diatoms and clear-water chironomids. We also found sedimentary accumulation of microplastic particles following the increase in Little Auk populations suggesting that seabirds potentially act as biovectors for plastic contamination. Our study demonstrates the diverse nature of climate-driven changes in the Arctic lacustrine environment with increased inorganic input from the more exposed catchment, larger nutrient delivery from the increased bird colonies at the surrounding mountain summits and subsequent alterations in aquatic communities.

Keywords Bird guano · Carbon · Chironomidae · Diatoms · Microplastic · Nitrogen

Introduction

Climate change alters biogeochemical cycling of major elements and nutrients, especially in regions with sparse vegetation, which are particularly sensitive to changes in surface energy and water balance (Zepp et al. 2007). This

phenomenon is most visible in the Polar Regions, where a recent pan-arctic greening of the tundra has been observed (Wookey et al. 2009). In addition to increasingly productive freshwater systems in the Arctic (Michelutti et al. 2005; Holmgren et al. 2010), in some regions, changes in the water balance and cryogenic processes have caused disappearance of lakes (Bouchard et al. 2013; Linderholm et al. 2018) as well as formation of new freshwater ecosystems known as permafrost thaw ponds (Vonk et al. 2015). Climate-driven biological reorganizations in the Arctic include increased primary production owing to longer summer growing seasons, increased algal habitat availability and enhanced catchment nutrient fluxes (Wrona et al. 2016). For example, melting permafrost will likely cause a release of nutrients into inlet streams driving changes in the ecosystem structure of lakes (Hobbie et al. 1999; Thienpont et al. 2013).

Another factor causing nutrient enrichment of Arctic lakes and their catchments is influence of climate on bird populations, as seabirds increasingly transport nutrients

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from the marine environment to their terrestrial nesting areas (Côté et al. 2010; Hargan et al. 2017), a phenomenon known as ornithogenic drainage (Smol 2016). Climate-mediated physical disturbances, such as changes in underwater light availability and thermal stability due to increased loading of organic or minerogenic matter from the terrestrial environment, may also lead to major disruptions to aquatic community structures (Vincent and Pienitz 1996; Nevalainen et al. 2015). In addition to enhanced biogeochemical cycling, greening terrestrial landscape and increasingly productive ecosystems (Forbes et al. 2010), accumulation of microplastic, especially in the oceans (Cole et al. 2011), is an increasing threat in the Arctic areas (Lusher et al. 2015). Besides the marine environment, microplastic particles are also transported to freshwater ecosystems, particularly in coastal areas, via the atmosphere or by biovectors, such as birds feeding in the ocean (Erkes-Medrano et al. 2015; Horton et al. 2017; Provencher et al. 2018). Microplastic particles cause threat to freshwater organisms through physiological problems (ingestion and digestion) and ecotoxicological effects (Dris et al. 2015). Although microplastics have been encountered even in remote arctic areas (Lusher et al. 2015), their distribution in High Arctic lakes is still mostly unknown.

Since observational records in the Arctic are scarce and short, indirect paleolimnological methods are required to reveal long-term environmental dynamics in high latitude lakes and their surroundings (Smol 2016). The paleolimnological record in surface and downcore lake sediments is based primarily on various physical and biogeochemical proxies and biological indicators, such as diatoms (Bacillariophyta) (Rühland et al. 2003; Rantala et al. 2017), chironomids (Diptera) (Quinlan et al. 2005; Luoto et al. 2019) and cladocerans (Crustacea) (Sweetman et al. 2008; Thienpont et al. 2015; Nevalainen et al. 2016). Physical proxies provide valuable lithological information, whereas biogeochemical proxies are particularly useful in tracking elemental cycling and lake-catchment interactions, such as bird effects using the sediment $\delta^{15}\text{N}$ signature (Stewart et al. 2013; Hargan et al. 2017). Diatoms are known to respond to pH and nutrient conditions (Tammelin et al. 2017; Pla-Rabés and Catalan 2018), chironomids to hypolimnetic oxygen and temperature (Quinlan and Smol 2002; Engels et al. 2014) and cladocerans to water quality and habitat changes (Jeppesen et al. 2011; Nevalainen 2012). The use of these paleolimnological proxies has enhanced understanding of the trajectories of climate-induced changes in northern aquatic ecosystems. Applying the multiproxy paleolimnological approach it is possible to assess long-term interactions between aquatic systems and their watersheds, i.e. lake-catchment coupling, and to find connections between different environmental realms, including the terrestrial, marine and atmospheric environments.

In this study, we analyzed fossil algal (diatoms) and invertebrate (chironomids, cladocerans and oribatid mites) communities together with physical (magnetic susceptibility, organic content), biogeochemical (carbon, nitrogen and their stable isotopes) and ecotoxicological (microplastics) proxies from a sediment profile collected from a High Arctic Lake Revvatnet in Svalbard (77°N). The aim was to build holistic understanding on long-term Arctic biogeochemical cycling and lake ecosystem shifts under the climate warming since the Little Ice Age. We hypothesize that long-term changes are climate-driven, but expect a complex interplay between varied environmental controls and multiple responses of the Arctic ecosystem. The study provides insights into linkages between the atmospheric, terrestrial and freshwater environments but also on the marine-derived influence, since extensive seabird colonies occupy the lake catchment area.

Study site

Lake Revvatnet (77.022°N, 15.368°E) is located in Hornsund, High Arctic Svalbard (Fig. 1a). Revvatnet is a glacial lake situated close to the Polish Polar Station Hornsund in an area characterized by pristine Arctic tundra and polar desert. The lake has infertile rocky shores, and a maximum depth of 26 m in the southern main basin (Fig. 1b). Based on epilimnetic limnological measurements at the end of July 2013, the water color was 0 PCU, pH 7.6, dissolved oxygen (DO) content 9.1 mg l⁻¹, specific conductivity 30 µg l⁻¹ and total dissolved solids (TDS) 10 µg l⁻¹. Arctic char (*Salvelinus alpinus*) were observed living in Revvatnet. Continuous water mixing occurs in Revvatnet during the summer, while it ceases during the winter when the lake freezes over (Nowiński and Wiśniewska-Wojtasik 2006). Revvatnet, which lies at an elevation of 30 m a.s.l. and has a surface area of about 0.9 km², is an overflow lake (Karczewski et al. 1981) with a network of streams and creeks entering from north and an outlet (Revelva) draining to Hornsund fjord (ocean bay) in the south (Fig. 1c).

The average present-day summer air temperature (June–August) in the area is 4.4 °C and the average annual precipitation is < 400 mm (Marsz and Styszyńska 2013). An increase in summer air temperature (~2 °C) since 1979 has been meteorologically observed (Marsz and Styszyńska 2013), but the biologically active vegetation period still lasts only ~2 months. The periglacial tundra catchment of Revvatnet lies on the Revbotnen and Revdalen post-glacial marine terraces between the Hornsund fjord and the mountain summits (Fig. 1c). The terrain consists of outwash plains and undulating ground moraine with sporadic marginal and lateral ridges, whereas hillsides feature solifluction lobes and talus cones. Barnacle geese (*Branta leucopsis*) are abundant in the adjacent Fuglebergsletta and extensive Little

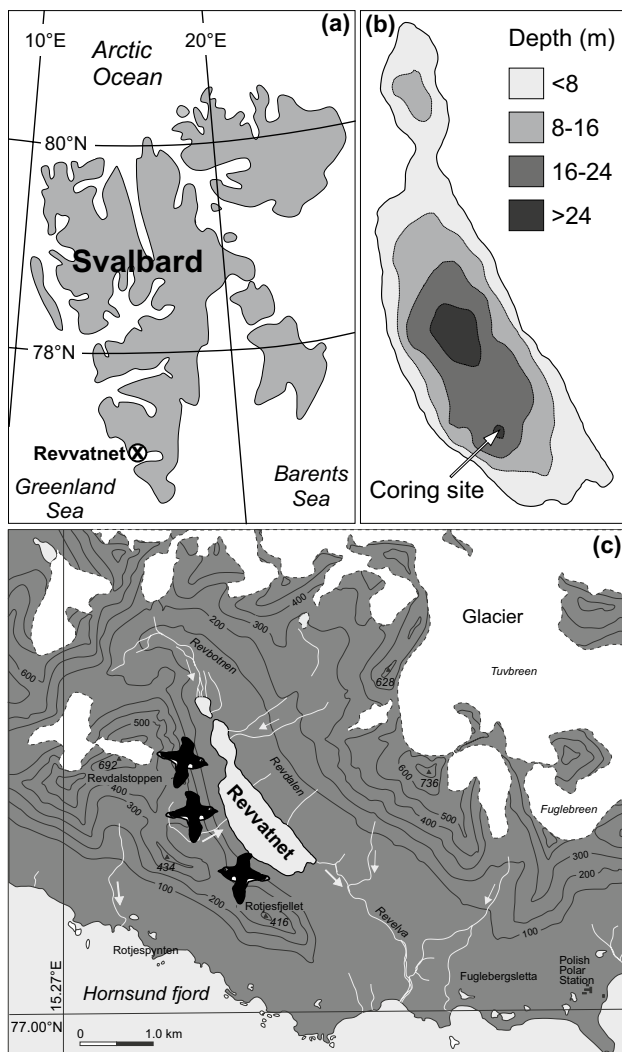


Fig. 1 The study site Revvatnet in Hornsund, Svalbard (a), bathymetric map and coring site (white arrow) (b) and catchment characteristics (c). Locations of Little Auk colonies are marked with bird symbols

Auk (*Alle alle*) colonies are present on the mountain slopes of Revdalstoppen and Rotjesfjellet, which drain into the lake. A recent expansion of bird colonies along the Hornsund coast has been observed (Wojczulanis-Jakubas et al. 2008; Zmudczyńska et al. 2009). The expansion of Little Auk colonies in the area appears to have begun during the early twentieth century (Gašiorowski and Sienkiewicz 2019).

Materials and methods

Sediments, chronology and sedimentological analyses

Of the collection of several sediment cores from Lake Revvatnet by Ojala et al. (2016), a 30-cm sediment profile RE2, taken from the southern part of the main basin, was used in the present study. We selected this specific core for the present study because of its distant location from the network of streams in the north (Revbotnen, Fig. 1) to avoid the dominant effect of stream sediments and to capture a variety of environmental changes and lake-catchment dynamics. The sampling was performed in June 2013 from a boat with a Kajak corer (Renberg 1991) and the sediments were subsampled at 0.5–1 cm intervals at the lake shore. Water depth at the coring site was 23.5 m.

For chronological control, ^{137}Cs analysis was performed at the Geological Survey of Finland using an EGandG Ortec ACE TM—2 K gamma spectrometer equipped with a four-inch NaI/Tl detector. The core was logged for magnetic susceptibility (Dearing 1999) with a Bartington MS2E1 surface-scanning sensor and subsampled for loss on ignition (LOI, + 550 °C for 2 h) at 1 cm resolution (Dean 1974).

Biogeochemical and microplastic analyses

Prior to the carbon analyses from sediment bulk organic matter, the fresh sediment was subjected to acid fumigation to remove carbonates, whereas nitrogen analyses were performed from natural sediment. Subsamples of 2–4 mg of freeze-dried and homogenized lake sediments were weighed and packed into tin capsules for elemental and stable isotope ($\delta^{13}\text{C}$, $\delta^{15}\text{N}$) composition of organic matter. 1-cm sample resolution was used in less compacted surface samples at 0–8 cm depth, whereas the lower part of the core was analyzed with 0.5 cm resolution. The analyses were performed with a FlashEA 1112 elemental analyser coupled with a Thermo Finnigan DELTA plus Advantage mass spectrometer. The results are expressed as delta values $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ (‰), described as $\delta = (R_{\text{sample}}/R_{\text{standard}} - 1) \times 1000$, where R equals $^{13}\text{C}/^{12}\text{C}$ and $^{15}\text{N}/^{14}\text{N}$, respectively. The reference standards are Vienna Pee Dee Belemnite for C and atmospheric N_2 for N. Proportions of organic C and total N (%) in organic matter were also used in the calculation of the $C_{\text{org}}/N_{\text{tot}}$ mass ratio that can be used as a source (allochthonous/autochthonous) indicator of organic matter (Meyers and Teranes 2001).

Topmost 10 cm were analyzed for microplastic particles using microscopic separation for identification and

classification (Karlsson et al. 2017). Plastic contamination from sampling and storage was avoided and controlled visually at identification. The separated size fractions included 100–500 μm , 500–1000 μm and > 1000 μm (Löder and Gerdt 2015). Therefore, since particles smaller than 100 μm were not analyzed, sample contamination from nanoplastics originating from e.g. clothing, sample storage and preparation was minimized. The results are expressed as number of plastic particles found per 1 cm^3 of dry sediment.

Diatom analysis

Samples for diatom analysis were prepared following standard procedures as described in Battarbee et al. (2001). The samples were analyzed using a 2-cm resolution. Organic matter was removed by oxidizing sediment samples with hydrogen peroxide (30% H_2O_2) followed by removal of carbonates with hydrochloric acid (37% HCl). Coarse mineralogenic matter was removed physically by swirling the sample solution in a beaker and decanting the diatom suspension. The sample residue was checked for absence of diatom valves prior to disposal. Samples were dried on coverslips and mounted with Naphrax, and a minimum of 300 diatom valves per sample were identified with a light microscope at 1000 \times magnification. Taxonomic determination was mainly based on the flora of Krammer and Lange-Bertalot (1986, 1988, 1991a, b), with nomenclature updated where relevant due to taxonomic refinements.

Chironomid and cladoceran analyses

Standard methods were applied to fossil chironomid analysis (Brooks et al. 2007). The samples were analyzed using a 1-cm resolution. The wet sediment was sieved through a mesh (100- μm) and the residue was examined under a stereomicroscope (25 \times magnification). Larval head capsules were extracted and mounted permanently with Euparal on microscope slides. Taxonomic identification following Brooks et al. (2007) was performed under a light microscope (400 \times magnification). The minimum chironomid head capsule number per sample was set to 50 (Heiri and Lotter 2001; Larocque 2001; Quinlan and Smol 2001). Alongside chironomid analysis, remains of cladocerans were picked and identified according to Szeroczyńska and Sarmaja-Korjonen (2007) and also oribatid mites were calculated following the procedure for environmentally extreme downcore sites (Luoto et al. 2013).

Statistical methods and data utilization

Hierarchical clustering was applied to separate stratigraphical diatom and chironomid zones. In the constrained cluster

analysis, we used the unweighted paired group method with arithmetic mean (UPGMA) as the algorithm and Bray–Curtis as the similarity index (dissimilarity threshold of 0.5 for a zone to be included). The clustering was carried out using the program Past3 (Hammer et al. 2001). Due to linear nature of the assemblage data, principal component analysis (PCA) was used to examine variation in diatom and chironomid communities. The species data were log₁₀ transformed prior to these analyses. The PCAs were carried out using the program Canoco 5 (Šmilauer and Lepš 2014). Diversity was assessed using the N2 effective number of occurrences (Hill 1973).

In addition to taxonomic assemblages, the diatom and chironomid data were examined for functional classification using ecological guilds and applying methodologies for algal (Rimet and Bouchez 2012) and macroinvertebrate (Schmera et al. 2017) functional ecology. The functional classification of diatoms was based on ecological guilds as delineated by Passy (2007) and Rimet and Bouchez (2012), including low profile taxa positioned at the bottom of the biofilm firmly attached to their substrate, high profile taxa extending to the upper layers of the biofilm (including colony forming diatoms), motile taxa capable of fast movement, and planktonic taxa. Each guild comprises taxa having developed diverse strategies to exploit resources and adapt to abiotic factors in a given environment, particularly with reference to nutrients, light, and physical disturbance. The chironomid feeding groups, including collector-gatherers and collector-filterers, were based on Merritt and Cummins (1996) and Mandaville (2002). As indicators for bird-impact, we used relative percentages of nitzschioid diatoms (Jones and Birks 2004; Keatley et al. 2009) and chironomids typical for lakes with significant bird influence in Svalbard, such as *Orthocladus trigonolabis*-type, *O. consobrinus*-type and *Metriocnemus eurynotus*-type (Brooks and Birks 2004; Luoto et al. 2016; Luoto and Ojala 2018).

To represent climate variability in Svalbard over the recent centuries, we used the data published by D'Andrea et al. (2012). Summer (June–August) temperature reconstructions are based on alkenone unsaturation in Lake Kongressvatnet, western Svalbard. To depict general trends, we used LOESS smooth with a span 0.2. The temperature data were obtained from the World Data Center for Paleoclimatology and NOAA's National Climatic Data Center, Paleoclimatology Branch website (<http://www.ncdc.noaa.gov/paleo/paleo.html>). The chronologies were matched using the ¹³⁷Cs peak in the Revvatnet RE2 sediment profile and further extrapolated deeper into the past.

Results

The ¹³⁷Cs stratigraphy of the present sediment core RE2 was found to be very similar with other ¹³⁷Cs stratigraphies in Svalbard (Appleby 2004; Chu et al. 2006; Luoto et al. 2015)

as well as other cores taken from Revvatnet and nearby Svartvatnet (Ojala et al. 2016). The ^{137}Cs activity in the RE2 core is well resolved with a single peak at the depth of 5.5–4.5 cm (Fig. 2). The peak can be linked with the atmospheric testing of nuclear weapons, with the onset of cesium fallout in the early 1950s and maximum fallout in 1963 CE. Therefore, the age horizon of ~1950 CE was assigned to

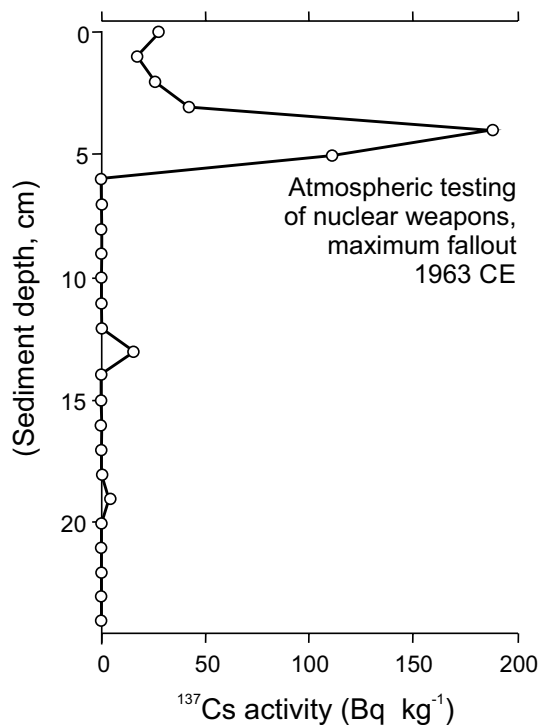


Fig. 2 ^{137}Cs activity in the sediment profile from Revvatnet, Svalbard

the sample at 5 cm. Chronological extrapolation provides an age estimate of ~1720 CE for the bottom core, but since the lower part of the sediment profile lacks chronological control, this estimate is uncertain. Considering the potential increase in recent sedimentation rates (Ojala et al. 2016), the extrapolated ages are more likely older than younger.

The physical and biogeochemical proxies showed rather consistent changes in the Revvatnet sediment profile (Fig. 3) and were partly linkable with changes in diatom and chironomid assemblages and summer temperature increase. In the initial part of the core (29–19 cm), magnetic susceptibility was low ($10\text{--}14 \text{ SI} \times 10^{-5}$) but began to increase towards the present ($> 20 \text{ SI} \times 10^{-5}$). A similar pattern was observed with $\delta^{13}\text{C}$ (from -29 to -25‰). In contrast, organic matter content (measured as LOI) was high in the initial phase (6–11%) but low between 18 and 0 cm (4–6%). Similar to organic matter, also total organic C (-0.7 to 1.4%), total N (0.1–0.2%) and $\text{C}_{\text{org}}/\text{N}_{\text{tot}}$ (-8.2 to 8.5) showed lower values in the upper sediment profile, with thresholds at 18 and 6 cm. $\delta^{15}\text{N}$ values had a deviating pattern showing a decreasing trend from the bottom core (from ~3 to 1‰) until a general increase in values from 10 cm onwards (mostly $> 3\text{‰}$).

From the sediment samples of Revvatnet, 89 diatom taxa were identified. The most abundant taxa included *Cyclotella rossii-comensis-tripartita* complex (mean abundance 18.7%, maximum abundance 48.0%), *Pseudostaurosira brevistriata* (10.6%, 21.9%) and *Achnanthisidium minutissimum* (7.0%, 20.3%). According to the cluster analysis, four diatom zones (I–IV) were separated (Fig. 4). In zone I between 28 and 16 cm, *C. rossii-comensis-tripartita* complex dominated and also *P. brevistriata* and *Stauroneis anceps* were common. In zone II between 14 and 10 cm, *P. brevistriata* became the most abundant taxon, while *C. rossii-comensis-tripartita*

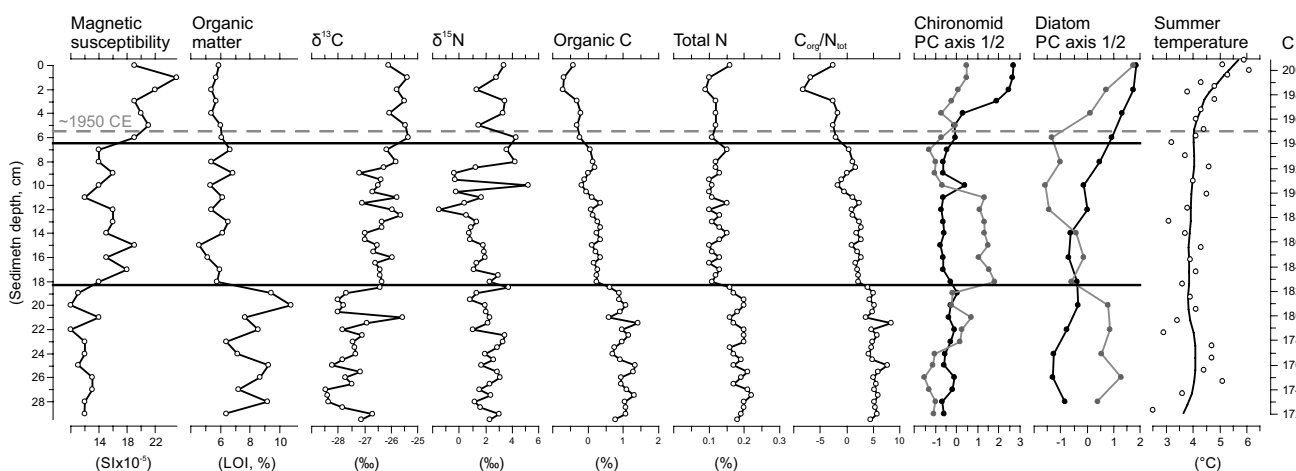


Fig. 3 Physical and biogeochemical sediment proxies compared with chironomid and diatom principal component (PC) axes 1 (black) and 2 (grey) scores. The temperature series using LOESS smoothing (span 0.2) is the sedimentary alkenone-based June–August air tem-

perature reconstruction from Svalbard (D'Andrea et al. 2012). The secondary axis is aligned with the ^{137}Cs horizon of the Revvatnet record (grey dashed horizontal line). Older extrapolated sediments are not reliably dated

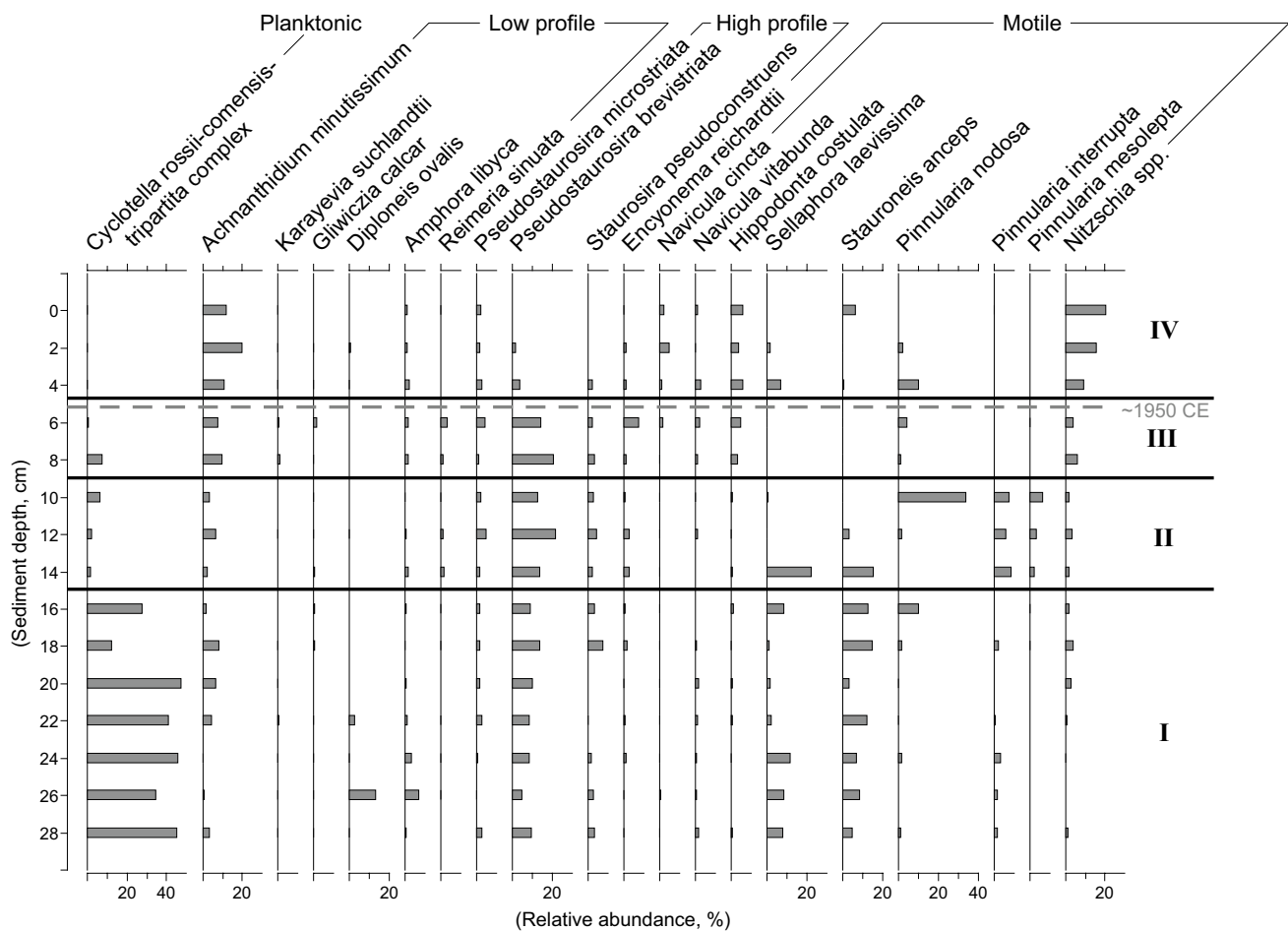


Fig. 4 Diatom stratigraphy of the most common taxa ($N \geq 10$, $\min \geq 5$) in Revvatnet grouped according to their ecological guilds. The floral zones (I–IV) were established using cluster analysis and the age horizon (grey dashed line) using ^{137}Cs analysis

complex markedly decreased. *P. brevistriata* continued to thrive in zone III between 8 and 6 cm, where taxa such as *A. minutissimum* and *Hippodonta costulata* also increased. Zone IV between 4 and 0 cm (from ~1950 CE until present) was dominated by *A. minutissimum* and *Nitzschia* spp. with simultaneous disappearances of the generally most abundant taxa *C. rossii-comensis-tripartita* complex and *P. brevistriata*.

The Revvatnet invertebrate stratigraphy consisted of 14 chironomid taxa, 1 cladoceran taxon and sporadic findings of oribatid mites. The most abundant chironomids included *Oliveridia tricornis* (mean abundance 50.7%, maximum abundance 90.4%), *Micropsectra radialis*-type (25.1%, 81.8%) and *Hydrobaenus lugubris*-type (11.6%, 38%). Similarly and almost concurrently with diatoms, four chironomid zones (I–IV) were separated (Fig. 5). Zone I between 29 and 19 cm was dominated by *O. tricornis* and *M. radialis*-type. In zone II between 18 and 11 cm, *M. radialis*-type disappeared and *O. tricornis* continued to dominate with *H. lugubris*-type. Zone III between 10 and 4 cm resembled

zone I, as *M. radialis*-type returned to the stratigraphy with high abundances. In the topmost zone I between 3 and 0 cm, previously predominant *O. tricornis* disappeared and *M. radialis*-type decreased. *Orthocladus trigonolabis*-type distinctly increased together with another member from the same genus, *O. consobrinus*-type. Similar to *O. tricornis*, the only cladoceran taxon, *Chydorus sphaericus*-type, disappeared permanently from the stratigraphy at 3 cm.

Due to relatively short gradient lengths in the diatom (2.0 SD) and chironomid (2.4 SD) data, linear ordination method (PCA) was used (Šmilauer and Lepš 2014). The first diatom PC axis ($\lambda_1 = 0.37$) explained 37.0% and the second axis ($\lambda_2 = 0.16$) 16.3% of all variance. The first chironomid PC axis ($\lambda_1 = 0.40$) explained 39.9% and the second axis ($\lambda_2 = 0.30$) 30.2% of the total variance. According to the primary axis scores, both diatoms and chironomids had negatives score in the initial part of the stratigraphy and high scores at the topmost samples (Fig. 6). In case of axis 2 scores, diatoms showed a decreasing trend from the bottom samples until the scores began to increase towards the

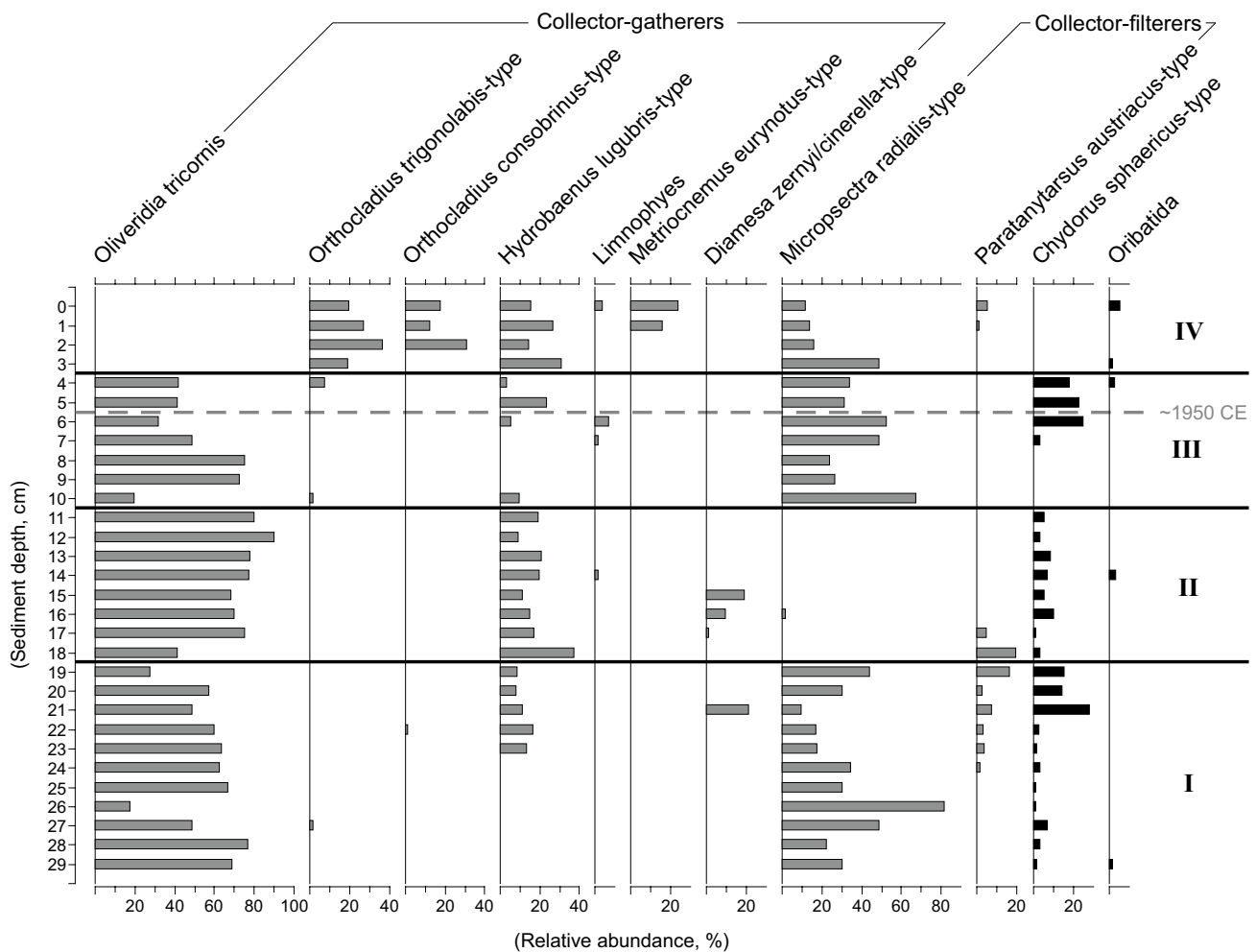


Fig. 5 Chironomid stratigraphy of the most common taxa ($N \geq 2$, $\min \geq 2$) in Revvatnet grouped according to their feeding guilds. The faunal zones (I–IV) were established using cluster analysis and

the age horizon (grey dashed line) using ^{137}Cs analysis. The relative abundances of cladoceran *Chydorus sphaericus*-type and oribatid mites are calculated from the total sum of invertebrates

present at 4 cm owing to distinct increases in *A. minutissimum* and *Nitzschia* spp. The chironomid axis 2 scores were low in the early phase of the stratigraphy (29–24 cm), after which they increased until a new decrease at 10 cm.

The functional classification of diatoms (Fig. 7) showed that planktonic taxa were abundant in the initial part of the sediment profile, between 28 and 16 cm, followed by a marked decrease that lasted until the present day. In contrast, low profile diatoms had their maximum abundances at the topmost samples between 8 and 0 cm, where high profile diatoms decreased. Motile diatoms were common throughout the stratigraphy but their highest abundances occurred between 14 and 0 cm. Nitzschioids showed a progressively increasing trend from the bottom of the core towards the present, with a significant shift at ~1950 CE when diatom taxonomic diversity (measured as N_2) also peaked.

Only two feeding guilds of chironomids were encountered from the stratigraphy (Fig. 8). Collector-gatherers

were present throughout the core, but collector-filterers were absent from the stratigraphy between 15 and 11 cm. Chironomids indicative of bird presence were absent in the initial phase of the sediment profile but showed moderate abundances (~10–20%) between 23 and 10 cm. After a short absence period, bird indicators reappeared at 7 cm and became highly abundant (> 80%) in the topmost samples (2–0 cm), where also the chironomid taxonomic diversity was highest.

Microplastic particles were found in the topmost sediment layers beginning from 2 cm (~1990s) upwards (Fig. 9). The highest microplastic accumulation (7.4 particles/cm³) was enumerated in the surface sample. Size fraction 100–500 μm was the most common in all samples where microplastics were encountered.

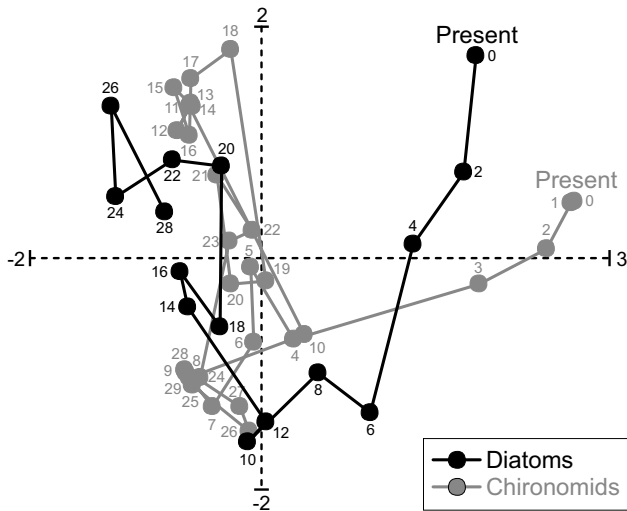


Fig. 6 Principal component analysis axis 1 and 2 scores for diatom (black) and chironomid (grey) samples of the Revvatnet sediment record

Discussion

Elemental cycling

Magnetic susceptibility increased in Revvatnet at two stages (Fig. 3), at 18 cm and at 6 cm, representing roughly the end of the Little Ice Age and the 1950s, respectively. Increases in magnetic mineral content of lake sediments are typically derived from more intense catchment erosion (Thompson et al. 1975; Dearing 1999; Ojala et al.

2017). In High Arctic environment with valley glaciers, magnetic susceptibility can also closely track changes in glacier oscillations through erosional effects (Nesje et al. 2001; Carlson et al. 2017). The recently increased values in Revvatnet correspond with the observed thinning rates in western Svalbard glaciers (Kohler et al. 2007) suggesting increasing melt water discharges (Fig. 1) and causing more intense erosion and transportation of mineral material delivery of melt waters into the Revvatnet basin from the northern inlets. The influence of glacier retreat, which began following the Little Ice Age, is also reflected in the marine sediment records from Hornsund fjord, but as the major glaciers fronts apparently retreated rapidly to the inner bays, the iceberg discharge to the fjord center became quickly limited (Pawłowska et al. 2016). In Lake Revvatnet, previous studies suggest a development towards a more turbid environment (Sienkiewicz et al. 2017). According to a contemporary survey (Ojala et al. 2016), the northern basin of Revvatnet (Fig. 1b) is significantly more turbid (11 formazin turbidity units, FTU) during summer open water season than the southern basin (4 FTU), from where the current core RE2 was collected. There is also a clear difference in the temperatures of the two basins, as the shallower northern basin is ~2 °C warmer. Despite the distinct difference in turbidity and temperature, no vertical stratification was observed suggesting continuous mixing during the summer. Since the Revvatnet basins are separated by a limnological and bathymetric sill, it inevitably has ecological significance.

The clear decrease in organic matter content at 18 cm (Fig. 3) is likely more related to increased transport of mineral matter from the catchment, as suggested by the magnetic

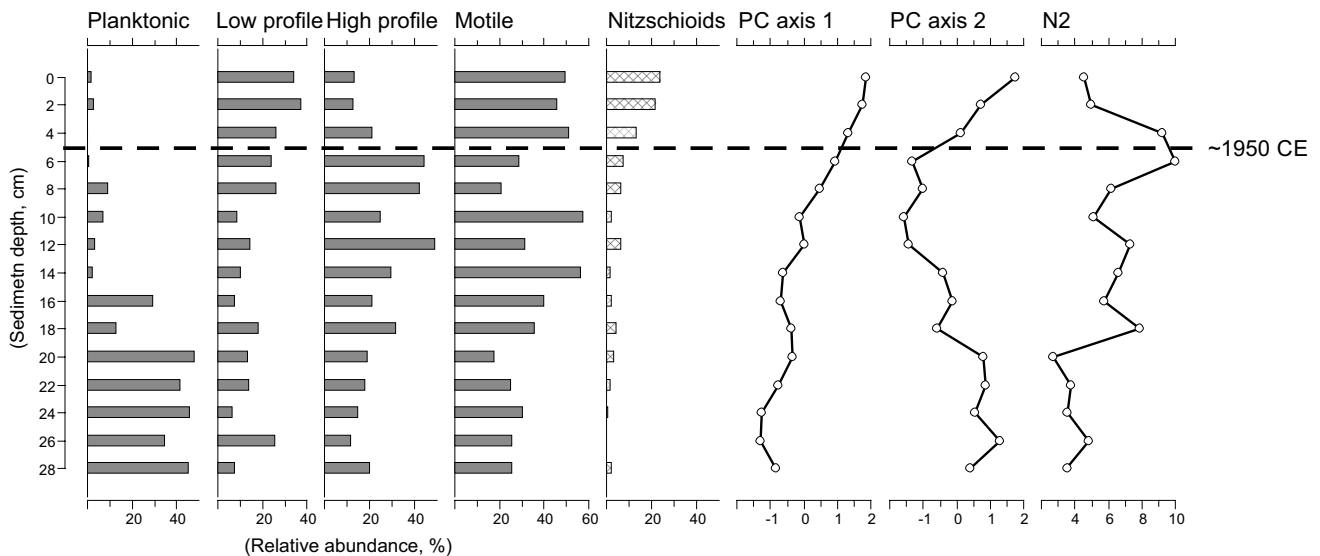


Fig. 7 Ecological guilds (grey) of diatoms (all diatoms included), relative share of nitzschioids (white pattern fill), principal component (PC) axis scores and effective diatom diversity (N2) in Revvatnet. The age horizon established using ¹³⁷Cs analysis is marked with a dashed line

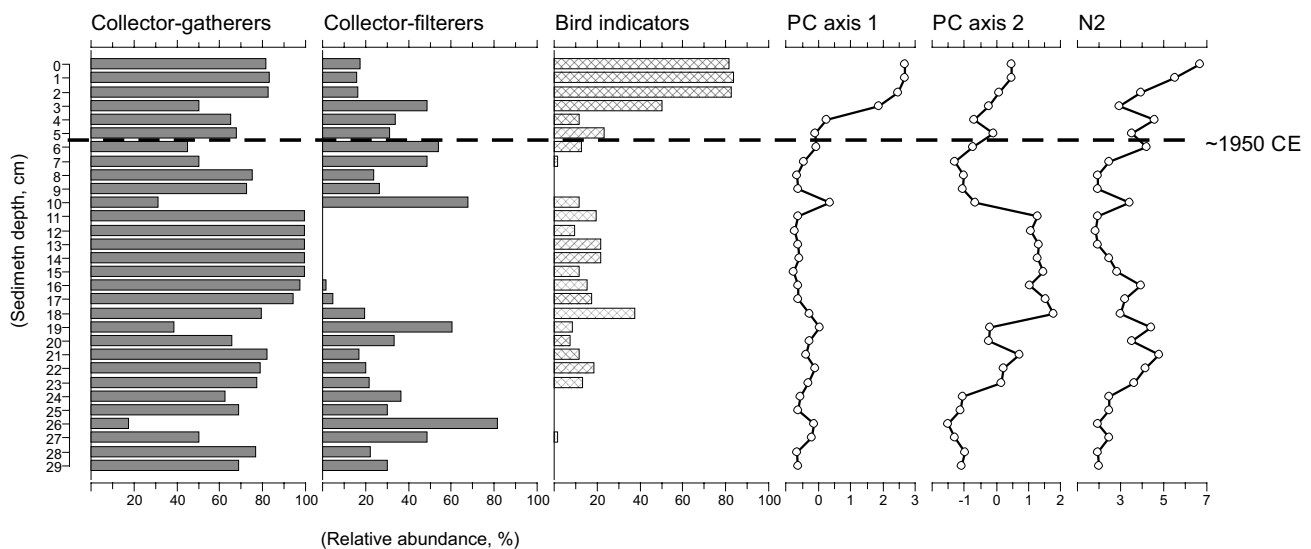


Fig. 8 Feeding guilds (grey) of chironomids (all chironomids included), relative share of taxa indicative of bird-impacted lakes (white pattern fill), principal component (PC) axis scores and effective

chironomid diversity (N2) in Revvatnet. The age horizon established using ^{137}Cs analysis is marked with a dashed line

susceptibility values, than a decrease in lake productivity (Meyers and Lallier-Vergès 1999). Nonetheless, this time horizon, corresponding roughly to the end of the Little Ice Age, was also a threshold for increasing $\delta^{13}\text{C}$ and also the total C and N contents had step shifts to lower relative values. These geochemical changes were accompanied by shifts in chironomid and diatom PC axis 2 values suggesting a secondary ecological response to the environmental change. The increased $\delta^{13}\text{C}$ values may indicate a change in benthic/periphytic production, which together with the other evidence, could reflect limnological and catchment processes that were driven by the end of the cold Little Ice Age and intensified melting of the glaciers, increased erosion and longer ice-free period (Guilizzoni et al. 2006). However, the exact timing and the changes in air temperature (D'Andrea et al. 2012) cannot be reliably connected with the current record due to the restrictions of the chronology. Therefore, the climate influence on the biogeochemical changes cannot be resolved from the data, although the majority of the long-term changes in the Arctic are considered climate driven (Smol and Douglas 2007b).

The second stage shift in environmental conditions occurred at 6 cm, corresponding to the mid-twentieth century (Fig. 3). The record high values in magnetic susceptibility during the upper part of the sediment profile and concurrent increase in air temperatures (D'Andrea et al. 2012) suggest further increase in glacier melt water delivery and catchment erosion. This interpretation is supported by the slightly decreased organic matter content. The $C_{\text{org}}/N_{\text{tot}}$ ratio showed a decreasing trend from the mid-twentieth century to the present indicating that productivity in Revvatnet shifted

from presumably mixed allochthonous and autochthonous to strictly autochthonous. A similar $C_{\text{org}}/N_{\text{tot}}$ shift has also been recorded from the adjacent geese-impacted pond Fugledammen (Luoto et al. 2015). Geese are present also in the catchment of Revvatnet, but Little Auks are particularly abundant (nesting cliff) fertilizing local tundra with their excrement through extensive transport of nutrients from the marine to terrestrial environment (Moe et al. 2009; Gąsiorowski and Sienkiewicz 2019). In addition to lowering $C_{\text{org}}/N_{\text{tot}}$ ratio, heavier $\delta^{15}\text{N}$ signature has been shown to be closely connected with seabird affected Arctic lakes (Griffiths et al. 2010; Stewart et al. 2013; Hargan et al. 2017). In Revvatnet, the $\delta^{15}\text{N}$ signal is rather variable during the latter half of the sediment profile compared to the earlier parts, but shows mostly ^{15}N -enriched values since the mid-twentieth century with initial peak already during the early twentieth century. Compared to $\delta^{15}\text{N}$ values measured from profundal lake sediments from Little Auk impacted sites in Greenland (~20‰) (González-Bergonzoni et al. 2017), the values in Revvatnet (3.4‰ in the surface sediment) remain rather low (Fig. 8), probably owing to its short residence time and smaller size of the colonies.

Ecological shifts

The diatom record in Revvatnet showed both gradual directional shifts and abrupt turnovers (Fig. 4). The most striking change was the sudden decrease in the centric *Cyclotella rossii-comensis-tripartita* complex, which dominated the diatom community during the early half of the stratigraphy (zone I). Although cosmopolitan by

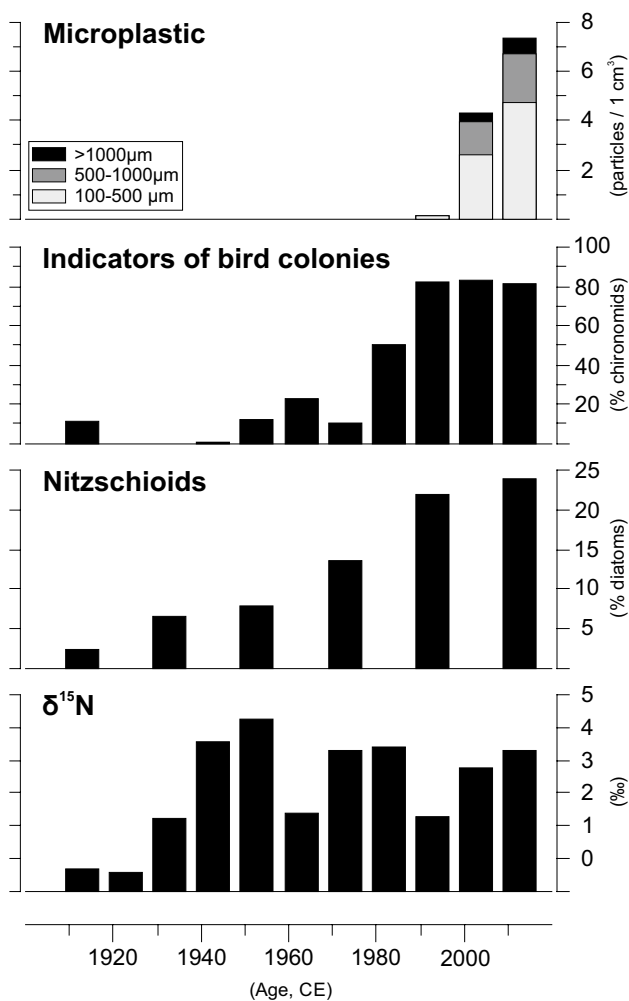


Fig. 9 Accumulation of microplastic particles with different size fractions in Revvatnet compared with indicators (chironomids, nitzschioid diatoms and $\delta^{15}\text{N}$) of seabird colonies (biovector)

nature, this species complex has been found a dominant component in the clear, oligotrophic lakes in East Greenland (Cremer and Wagner 2004) and indicates stability of surrounding Arctic landscape (Perren et al. 2012). Small cyclotelloids are generally considered sensitive to changes in water clarity and physical structure of the water column, while changing nutrient regimes may also govern their abundance (Rühland et al. 2015; Saros and Anderson 2015). While the *C. rossii-comensis-tripartita* complex rapidly decreased at 14 cm (zone II) and finally disappeared at 4 cm (zone IV), *Achnanthydium minutissimum*, a benthic pennate species, increased towards the present, especially from 8 cm (zone III) onward. In warmer climates *A. minutissimum* is known as a primary colonizer in disturbed environments with turbid waters and watersheds characterized by loss in vegetation cover (Peterson and Stevenson 1992; Caballero et al. 2006). Therefore, these shifts in the diatom community of Revvatnet may

indicate a transition from a clear and oligotrophic lake toward a more turbid environment with diminished light penetration.

In all, the diatom assemblages in Revvatnet appear to be more responsive to catchment disturbances and prolonged delivery of suspended fine-grained mineral matter rather than climate warming directly. An increase of magnetic susceptibility in upper sediment agrees well with this observation (Fig. 3). However, the distinct decrease in *Pseudostaurastira brevistriata* at 4 cm (zone IV), representing the 1960s, can be linked with climate change, as the species has been observed to decline in cold, oligotrophic, deep lakes with extensive ice cover due to recent climate warming (Rühland et al. 2003). Also the increase in *A. minutissimum* has been related to climate warming in several Arctic records (Antoniades et al. 2004; Keatley et al. 2006; Lim et al. 2008; Paul et al. 2010), but it is also found in mossy habitats (Griffiths et al. 2017) suggesting that the lake-catchment area has become more productive. The recent increase in *Nitzschia* spp. is also likely related to increase in nutrients, as previously recorded from the Canadian Arctic (Michelutti et al. 2007). In all, the recent changes in diatom assemblages are likely related directly (e.g. longer ice-free period) or indirectly (increased turbidity and nutrients) to climate change.

Chironomids respond primarily to air or water temperature (Eggermont and Heiri 2012), but the taxonomic composition of Revvatnet consists thoroughly of taxa with the coldest temperature optima (Heiri et al. 2011). This is not surprising, since the study lake is situated in the High Arctic and is fed by glaciers. The chironomid fauna resembles that found in Einstaken (Luoto et al. 2011), a lake located in northernmost Svalbard (80°N) with a mean July air temperature of only ~2 °C. However, compared to an adjacent (15 km) lake Svartvatnet, located on the southern side of the Hornsund fjord, the chironomid compositions differ despite the similar deep oligotrophic character of the two lakes. For example, *Micropsectra contracta*-type, which has dominated in Svartvatnet for at least the last 5000 years (Luoto et al. 2018), is completely absent in the Revvatnet record (Fig. 5). Since *M. contracta*-type has a warmer temperature preference (Heiri et al. 2011) than the taxa encountered in Revvatnet, it is likely that its absence from the present record is caused by a continuous production-season inflow of cold water (glacial melting influence), which is not a factor in Svartvatnet.

The greatest shifts in the chironomid communities in the RE2 record occur when *M. radialis*-type disappears between 18 and 11 cm (zone II) and when *Oliveridia tricornis* vanish from the stratigraphy at 3 cm (zone IV) (Fig. 5). In Einstaken, decrease in *M. radialis*-type was related to climate warming (Luoto et al. 2011), but this is not the case in Revvatnet since its absence occurs when the temperatures were still very low following the Little Ice

Age (D'Andrea et al. 2012), and furthermore, it reappeared when the climate warmed during the twentieth century. Since the timing when *M. radialis*-type disappeared coincides roughly with changes in diatoms and when *Cyclotella rossii-comensis-tripartita* complex began to decrease (Fig. 4), it may be that it is related to changes in minerogenic water turbidity. Nonetheless, since the changes in chironomids and diatoms were not fully synchronous at this time, it may also be that *M. radialis*-type was responding to reduced hypolimnetic oxygen conditions caused by the increased turbidity and subsequent potential vertical stratification. While it is possible that the study lake thermally stratified in the past (at the end of the Little Ice Age) by enhanced glacial input, cold glacial meltwater may also have disrupted the thermal stratification of the water column when climate began to warm in the twentieth century, thereby altering the diatom (i.e. shift from planktonic to benthic) and chironomid assemblages (i.e. shift in the oxy-stressor *M. radialis*-type). This interpretation would be similar to a previous diatom study from Revvatnet (Sienkiewicz et al. 2017). Also the extirpation of *O. tricornis* can be related to changes in water quality, because it corresponds with the sample where also the only cladoceran in the record, *Chydorus sphaericus*-type, disappears. In another site located in Hornsund, Fugledammen, the decrease in *C. sphaericus*-type towards the present was linked with increased nutrient conditions that turned the pond murky (Luoto et al. 2015). Therefore, the chironomid and cladoceran shift at 3 cm can also be related to limno-optical changes in the lake environment. These changes were likely catchment originated, as indicated by the appearance of *Metriocnemus eurynotus*-type, which is a semiterrestrial taxon (Brooks et al. 2007), suggesting material transport from the watershed. Nonetheless, similar to the recent increase in *Nitzshia* spp. (Fig. 4), the chironomid zone I is characterized by appearance of *Orthocladus* species that also clearly indicate increased trophic state (Brooks and Birks 2004).

The algal and invertebrate communities, representing primary and secondary producers, have generally parallel community dynamics in Revvatnet (Figs. 4, 5), which is also shown by the ordination results suggesting a shift that is directionally similar (Fig. 6). Interestingly, both ordinations clearly separate the recent samples into their own primary axis cluster showing that the communities of the recent decades are unprecedented compared to the earlier biostratigraphy. This finding supports previous evidence that significant changes in Arctic aquatic communities are currently occurring, and that critical ecological thresholds are being crossed (Smol and Douglas 2007a, b; Axford et al. 2009). The current taxonomic records also suggest that the climate influence may not always be direct but the recent changes may also occur due to indirect climate influence,

i.e. increase in the bird colony, hence suggesting multiple community responses (Smol 2010).

In addition to taxonomic compositions, applications of functional characteristics of aquatic organisms can be valuable for understanding ecosystem processes and dynamics (Jeppesen et al. 2001; Nevalainen and Luoto 2017; Kivilä et al. 2019). In Revvatnet, the changes in diatom ecological guilds support the environmental evidence derived from taxonomic compositions. A decrease in planktonic diatoms at the core depth of 14 cm and increases in low profile (such as achnantheoids) as well as motile (such as nitzschoids and naviculoids) life forms (Fig. 7), which include diatoms that typically tolerate physical disturbances (Tapolczai et al. 2016), provide uniform evidence for increased water turbidity. Only two chironomid feeding guilds were encountered in the Revvatnet sediment profile (Fig. 8), which hampers more detailed interpretation.

As birds nesting or grazing in tundra are known to influence Arctic lakes (Mariash et al. 2018), we separated diatom and chironomid taxa that are typically associated with lakes that have significant bird populations in their catchment (e.g. Brooks and Birks 2004; Jones and Birks 2004). As a result, nitzschoid diatoms and chironomid bird indicators both showed an increasing trend from ~24 cm onward with largest shifts during the most recent decades (Figs. 7, 8). In the previous study from the neighboring pond Fugledammen located in the Fuglebergsletta (Fig. 1c), it was shown that the impact of Barnacle geese significantly increased during the twentieth century (Luoto et al. 2015). Similar evidence of geese population growth has also been found from northeastern Svalbard (Luoto et al. 2014). A study of peat sequences from the current study areas also suggested major Little Auk population growth, with a threshold occurring at ~1920 CE (Gąsiorowski and Sienkiewicz 2019). Therefore, it appears to be clear that the bird impact has become a significant factor in Svalbard over the past decades, consequently altering the aquatic ecosystems and watersheds. The response of diatoms to nutrients can be direct but in case of chironomids the bird impact may derive from deteriorated oxygen conditions (Stewart et al. 2013). However, there are no signs of current summertime vertical stratification in Revvatnet (Ojala et al. 2016), reducing the likelihood that oxygen decrease is driving the increase in the bird indicator taxa. In addition, the presence of *Micropsectra* suggests well-oxygenated hypolimnion, since it is an “oxy-stressor”, a genus that requires high oxygen levels (Brodersen et al. 2008). An option is that the chironomid response is due to habitat change, since the bird indicators *O. trigonolabis*-type and *O. consobrinus*-type prefer more productive lakes in Svalbard with aquatic macrophytes present (Brooks and Birks 2004; Luoto et al. 2016). An unpublished analysis of a sediment core from the northern basin (RE8 in Ojala et al. 2016) showed that while other invertebrate remains were

very scarce, ostracods (Podocopida) were numerous, unlike in the southern basin where they were completely absent. Previous studies from Hornsund (Luoto et al. 2015) and elsewhere in Svalbard (Luoto et al. 2014) have shown that ostracods increase under a bird-induced eutrophication process and they are more common in contemporary samples of moderately impacted (nutrient-rich) lakes of Fuglebergsletta with dense macrophyte growth (Luoto et al. 2016), hence supporting the interpretation on habitat change.

There are also major changes in taxonomic diversity in the RE2 record during the recent decades. In diatoms, the diversity increased until ~1950 s but then rapidly decreased (Fig. 7). Chironomid diversity also increased towards the 1950s, but unlike with diatoms, it remained elevated and further increased during the most recent decades (Fig. 8). The recent decline in diatom diversity is likely caused by the increased turbidity restricting many planktonic and pennate diatoms (Bradshaw et al. 2000). In fact, it has been shown that concurrently increasing water turbidity and nutrient levels can cause decline in planktonic and benthic diatom diversities, while chironomid diversity simultaneously increased (Luoto et al. 2017) supporting the interpretation of turbidity and nutrient control on the recent ecosystem changes in Revvatnet.

Microplastic delivery

A scan for microplastic accumulation in the Revvatnet RE2 sediments revealed that the first encountered particles appeared at 2 cm depth, representing the 1990s, with an increasing abundance towards the present (Fig. 9). Microplastics are harmful for freshwater organisms, especially since their bioaccumulation potential increases with decreasing size (Lee et al. 2013). In the Revvatnet sediment, the most common particle size was the smallest analyzed size fraction (100–500 μm) that is potentially most harmful for chironomids, since they can swallow these sized particles with their mouthparts, but have great risk of causing an obstruction (Scherer et al. 2017). In the samples, where microplastics were found, collector-filterers decreased (Fig. 4), which might suggest a causal relationship. Also the fine-mesh filter-feeding cladoceran *Chydorus sphaericus*-type disappeared from the topmost samples, this change possibly being related to microplastic fragment consumption. However, since microplastic occurrences are represented only in three uppermost and most recently deposited sediment samples (2–0 cm) and no ecotoxicological experiments have been performed, these interpretations remain merely speculative.

Though mostly investigated in marine environments, microplastics are also transported to Arctic lakes directly by humans or via the atmosphere and biovectors, such as seabirds (Wagner et al. 2014), which feed in the ocean

(ingestion) and defecate in their terrestrial nesting grounds. Significant amounts of microplastic have been found in several arctic seabirds, including Little Auks (O'Hanlon et al. 2017). In fact, Little Auks, which are abundant in the catchment of Revvatnet, are particularly susceptible to ingesting microplastic, since they predominantly feed on smaller prey items (such as copepods) and therefore are more likely to mistake microplastic for prey, or ingest it accidentally whilst foraging (Amélineau et al. 2016). In the current record, the chironomid-, diatom- and $\delta^{15}\text{N}$ -inferred increase in Little Auk colonies initiate just before the appearance of microplastics (Fig. 9) indicating availability of an efficient biovector for marine derived contaminants. In addition to the physiological damaging effects resulting from direct ingestion of microplastics by aquatic organisms, microplastics also act as vectors for organic pollutants (e.g. other POPs) (Bakir et al. 2014). Although this prospective study cannot resolve the response of aquatic taxa to microplastic pollution, our findings suggest high probability for Arctic freshwater biota to encounter microplastics and a potential for trophic interactions and functional changes. However, further research is required to understand the particle transport and effects of microplastic-biota interactions within increasingly productive High Arctic freshwater environments.

Conclusions

The present study showed major changes in the physical, biogeochemical and ecological environment of the High Arctic Revvatnet and its catchment. The environmental changes occurred at two stages during the end of the Little Ice Age and the mid-twentieth century. Our results suggest progressively increasing transport of mineral matter from the catchment that originated from glacial melt water erosion, and subsequently caused significantly more turbid waters in Revvatnet. We recorded increased biogeochemical cycling also through a more productive catchment and higher lake autochthony. Based on the increase in nitzschoid diatoms, bird indicating chironomids and decline in the $C_{\text{org}}/N_{\text{tot}}$ ratio, for example, the influence of expanding Little Auk colonies in the catchment was noticeable but superimposed with the other climate warming impacts on Revvatnet. The increased Little Auk population has also produced a pathway for microplastic delivery, as we reported microplastic particles in the sediments since the 1990s with a progressive accumulation rate. Overall, we recorded multiple ecological responses and functional shifts of the aquatic ecosystem, whilst the influences of climate change as the ultimate driving force are undisputable.

As also demonstrated by the current study, climate-forced changes cause cascading effects on the terrestrial and aquatic environments in the Arctic, with increased glacier

melt and erosion, more productive catchments and limnologically altered freshwater basins. Based on the present results, it becomes clear that the patterns in biogeochemical elemental cycling have changed with significant thresholds being crossed. The driving forces of the major changes of Arctic lakes are evidently anthropogenic through human-induced climate warming and delivery of pollutants.

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