

This is a self-archived version of an original article. This version may differ from the original in pagination and typographic details.

Author(s): Saine, Sonja; Aakala, Tuomas; Purhonen, Jenna; Launis, Annina; Tuovila, Hanna; Kosonen, Timo; Halme, Panu

Title: Effects of local forest continuity on the diversity of fungi on standing dead pines

Year: 2018

Version: Accepted version (Final draft)

Copyright: © 2017 Elsevier B.V.

Rights: CC BY-NC-ND 4.0; CC BY-NC-ND 4.0

Rights url: https://creativecommons.org/licenses/by-nc-nd/4.0/; https://creativecommons.org/

Please cite the original version:

Saine, S., Aakala, T., Purhonen, J., Launis, A., Tuovila, H., Kosonen, T., & Halme, P. (2018). Effects of local forest continuity on the diversity of fungi on standing dead pines. Forest Ecology and Management, 409, 757-765. https://doi.org/10.1016/j.foreco.2017.11.045

EFFECTS OF LOCAL FOREST CONTINUITY ON THE DIVERSITY OF FUNGI ON STANDING

2 **DEAD PINES**

- 3 Saine Sonja, Aakala Tuomas, Purhonen Jenna, Launis Annina, Tuovila Hanna, Kosonen Timo & Halme
- 4 Panu

1

5 APPENDIX B. Responses of individual species

- 6 In the main text of this article we report the results of the community analyses. Here we report analyses on
- 7 the responses of single species analyses using the same explanatory variables. We also briefly report the
- 8 methods of the single species analyses and shortly discuss the results.

9 **METHODS**

Responses of single species were analyzed with a Mixed Effects Logistic Regression (n = 52). The aim was to study which environmental variables explain occurrences of each species the best. Species that occurred on ≥ 10 study trunks were included into the analysis (n = 14). A Mixed Effects Logistic Regression with a binomial distribution and a log-linear link function was conducted separately for each species. Explanatory variables were the same as in the GLMM (see the main text). Site and trunk identities were included into the model as hierarchically structured random effects by nesting trunks within sites. The model selection was

conducted as in the GLMM. The analysis was performed in R (version 3.3.2; R Core Team, 2016) using

17 function "glmer" from the package "lme4" (Bates et al., 2016).

RESULTS

16

18

- 19 Studied variables explained the presence of four species altogether. For the rest, the final models did not
- 20 include any statistically significant variables. Occurrences of Actidium hysterioides were negatively affected
- 21 by years from death (Table B.1). The final model also included dead wood diversity that had a marginally
- 22 significant negative effect on the species (Table B.1). Occurrences of *Chaenothecopsis pusiola* were best
- 23 explained by a negative effect of dead wood diversity, a negative effect of age at death, and a positive effect
- of years from death (Table B.1). These were all variables included in the final model. Canopy openness
- 25 appeared to have a positive effect on the occurrences of *Glonium nitidum* (Table B.1). Number of stumps
- and years from death were the other variables included in the final model (Table B.1). The negative effect of

years from death was marginally significant (Table B.1). Occurrences of *Micarea elachista* were best explained by a negative effect of dead wood diversity (Table B.1). Another variable included in the final model was the number of stumps that had a marginally significant positive effect on the species as well (Table B.1).

Species final models of which included marginally significant effects of certain variables were *Micarea prasina* (a positive effect of years from death), *Mycocalicium subtile "thin"* (a negative effect of management intensity), Pyrenomycete sp. 4 (a negative effect of dead wood diversity, and a positive effect of canopy openness), and *Trichaptum fuscoviolaceum* (a negative effect of years from death) (Table B.1).

Table B.1. Results from the Mixed Effects Logistic Regression for individual species (n = 52 for each). Cells show estimates (B), standard errors (SE), z values, and statistical significances (P). Dead wood (DW) diversity was calculated with Shannon's diversity index. The units used for variables are in brackets. Abbreviations: YFD = years from death, AAD = age at death, stumps = management intensity, canopy = canopy openness.

Species		В	SE	z value	P
Actidium hysterioides	(Intercept)	-1.72	0.76	-2.27	0.023
	DW diversity	-1.09	0.62	-1.75	0.081
	YFD (y)	-1.42	0.69	-2.07	0.039
Chaenothecopsis pusiola	(Intercept)	-0.75	0.35	-2.16	0.031
	DW diversity	-0.76	0.37	-2.05	0.041
	AAD (y)	-0.81	0.41	-1.97	0.049
	YFD (y)	0.97	0.41	2.38	0.017
Glonium nitidum	(Intercept)	0.18	0.61	0.30	0.763
	Stumps (pc ha ⁻¹)	1.00	0.67	1.48	0.138
	YFD (y)	-1.04	0.55	-1.90	0.058
	Canopy (%)	1.75	0.78	2.25	0.025
Micarea contexta	(Intercept)	-1.73	0.69	-2.51	0.012
	YFD (y)	0.55	0.44	1.25	0.213
Micarea elachista	(Intercept)	-3.19	0.92	-3.46	< 0.001
	DW diversity	-2.58	0.92	-2.81	0.005
	Stumps (pc ha ⁻¹)	0.84	0.47	1.80	0.072
Micarea melaena	(Intercept)	0.81	0.46	1.77	0.076
	AAD (y)	-0.43	0.43	-1.01	0.315
Micarea misella	(Intercept)	-0.59	0.41	-1.44	0.149
	YFD (y)	0.12	0.34	0.35	0.728
Micarea prasina	(Interept)	-1.20	0.71	-1.69	0.091
	YFD (y)	0.74	0.44	1.68	0.093
Mycocalicium subtile "big"	(Intercept)	-0.86	0.32	-2.74	0.006
	AAD (y)	-0.51	0.36	-1.44	0.150
Mycocalicium subtile "thin"	(Intercept)	-2.34	1.15	-2.03	0.043
	Stumps (pc ha ⁻¹)	-1.70	0.99	-1.72	0.085
Pyrenomycete	(Intercept)	-1.31	0.34	-3.82	< 0.001
sp. 1	Canopy (%)	0.33	0.39	0.85	0.398
Pyrenomycete sp. 2	(Intercept)	-2.46	2.55	-0.96	0.335
	DW diversity	-1.15	1.28	-0.90	0.370
	AAD (y)	-1.25	1.33	-0.94	0.346
	YFD (y)	-1.04	1.21	-0.86	0.388
Pyrenomycete sp. 4	(Intercept)	-0.65	0.44	-1.49	0.136
	DW diversity	-0.82	0.43	-1.94	0.053
	Canopy (%)	0.99	0.51	1.93	0.053
Trichaptum	(Intercept)	-1.71	0.61	-2.83	0.005
fuscoviolaceum	YFD (y)	-0.84	0.49	-1.70	0.089

DISCUSSION

Occurrences of *Chaenothecopsis pusiola* showed a positive association with time since tree death, and for *Micarea prasina* the positive effect was nearly statistically significant. The species occurred more likely on trunks that had died longer time ago. The positive effect of time since tree death on *C. pusiola* might be explained by its suggested parasitic relationship with lichens and non-symbiotic algal colonies (Tuovila, 2013). Also, more suitable habitats form with time, as the species prefers decorticated wood (Lõhmus and Lõhmus, 2001). *M. prasina* is a crustose lichen especially common in old-growth forests (Stenroos et al., 2015). Presumably, this slow-growing species (Stenroos et al., 2011) benefits from long periods since tree death like lichens in general.

Years from death had a negative effect on pyrenomycetes *Actidium hysterioides* and *Glonium nitidum* and polypore *Trichaptum fuscoviolaceum*, yet the effect was not statistically significant for the latter two. All these species might be early successional species. Many pyrenomycetes latent in the wood are abundant in initial decay stages (Heilmann-Clausen, 2001; Hendry et al., 2002). Additionally, increasing moisture content with proceeding decomposition (Sollins et al., 1987) might hinder these species adapted to dry conditions (Boddy et al., 1989, 1985). *T. fuscoviolaceum* is a pioneer species that is often among the initial decomposers (Niemelä et al., 1995; Renvall, 1995). The species loses in competitive ability or due to depleting resources when late-stage specialists colonize the community (Rayner and Boddy, 1988; Stokland et al., 2012).

C. pusiola responded negatively to the increasing trunk age at death. The species seems to occur frequently on decorticated, decayed surfaces on the base of the boles (Hanna Tuovila, personal communication). Such microhabitat patches might be more common in younger trunks due to the differences in decay succession. Old standing kelo trees, for example, rarely offer such microhabitats. Therefore, the species might prefer trunks that have died at younger age.

Dead wood diversity had a negative effect on *C. pusiola, Micarea elachista, A. hysterioides* and Pyrenomycete sp. 4, although the effect was not statistically significant for the latter two. *C. pusiola* also occurs in managed forests as long as suitable substrates are available (Hanna Tuovila, personal communication), and therefore the species might not be dependent on old-growth forests *per se*. The

negative effect on *M. elachista* might indicate that the species is not dead wood dependent as such since it can also grow on old living trees (Coppins, 1983; Czarnota, 2007). Like many pyrenomycetes, also *A. hysterioides* and Pyrenomycete sp. 4 might be associated with early stages of decomposition (Heilmann-Clausen, 2001; Hendry et al., 2002). The emergence of more specialized species with increasing dead wood diversity might have an adverse effect on these species.

Pyrenomycetes *G. nitidum Ellis* and Pyrenomycete sp. 4 showed a positive response to canopy openness, yet the effect was not statistically significant for the latter. Pyrenomycetes in general are characterized by high resistance to water stress (Boddy et al., 1989, 1985). Consequently, the competitive superiority of these species in dry circumstances might explain the result.

Altogether, negative responses to local continuity were predominant among individual species. Such responses could be expected from generalists that lose to late-stage specialists in competitiveness (Marvier et al., 2004). Kruys et al. (1999) hypothesized that species dependent on dead wood continuity require habitats that are scarce within a landscape. Presumably, these species are specialists and rare. In our study, such species did not probably have enough occurrences to be included in the analyses. Some of these species might be *e.g.* veteran tree specialists that inhabit the oldest trunks for which the age parameters were not successfully quantified. Therefore, more research on rare and specialized species is required to clarify their relationship with local continuity.

85 **REFERENCES**

- 86 Bates, D., Maechler, M., Bolker, B., Walker, S., Christensen, R.H.B., Singmann, H., Dai, B., Grothendieck,
- 87 G., Green, P., 2016. lme4: Linear Mixed-Effects Model using 'Eigen' and S4, version 1.1-12.
- https://cran.r-project.org/web/packages/lme4/lme4.pdf. Accessed 30.11. 2016.
- Boddy, L., Gibbon, O.M., Grundy, M.A., 1985. Ecology of Daldinia concentrica: effect of abiotic variables on mycelial extension and interspecific interactions. Trans. Br. Mycol. Soc. 85, 201–211.
- 91 doi:10.1016/S0007-1536(85)80183-2
- 92 Boddy, L., Owens, E.M., Chapela, I.H., 1989. Small scale variation in decay rate within logs one year after
- 93 felling: effect of fungal community structure and moisture content. FEMS Microbiol. Ecol. 62, 173-
- 94 184. doi:10.1016/0378-1097(89)90110-9
- Coppins, B.J., 1983. A taxonomic study of the lichen genus Micarea in Europe. Bull. Br. Museum (Natural Hist. Bot. 11, 1–204.
- Czarnota, P., 2007. The lichen genus Micarea (Iecanorales, Ascomycota) in Poland. Polish Bot. Stud. 23, 1–98 197.
- Heilmann-Clausen, J., 2001. A gradient analysis of communities of macrofungi and slime moulds on decaying beech logs. Mycol. Res. 105, 575–596. doi:10.1017/S0953756201003665
- Hendry, S.J., Boddy, L., Lonsdale, D., 2002. Abiotic variables effect differential expression of latent infections in beech (Fagus sylvatica). New Phytol. 155, 449–460.
- Kruys, N., Fries, C., Jonsson, B.G., Lämås, T., Ståhl, G., 1999. Wood-inhabiting cryptogams on dead Norway spruce (Picea abies) trees in managed Swedish boreal forests. Can. J. For. Res. 29, 178–186.
- 105 doi:10.1139/x98-191
- 106 Lõhmus, P., Lõhmus, A., 2001. Snags, and their lichen flora in old Estonian peatland forests. Ann. Bot. 107 Fenn. 38, 265–280.
- Marvier, M., Kareiva, P., Neubert, M.G., 2004. Habitat destruction, fragmentation, and disturbance promote
- invasion by habitat generalists in a multispecies metapopulation. Risk Anal. 24, 869–878.
- 110 doi:10.1111/j.0272-4332.2004.00485.x
- Niemelä, T., Wallenius, T., Kotiranta, H., 1995. Interactions of fungi at late stages of wood decomposition.
- 112 Ann. Bot. Fenn. 32, 141–152.
- Rayner, A.D.M., Boddy, L., 1988. Fungal decomposition of wood: its biology and ecology. John Wiley &
- Sons, Chichester.
- 115 R Core Team. 2016. R: A language and environment for statistical computing. Version 3.3.2. R Foundation
- for Statistical Computing, Vienna, Austria. https://www.r-project.org/. Accessed 30.11. 2016.
- 117 Renvall, P., 1995. Community structure and dynamics of wood-rotting Basidiomycetes on decomposing conifer trunks in northern Finland. Karstenia 35, 1–51.
- Sollins, P., Cline, S.P., Verhoeven, T., Sachs, D., Spycher, G., 1987. Patterns of log decay in old-growth Douglas-fir forests. Can. J. For. Res. 17, 1585–1595. doi:10.1139/x87-243
- 121 Stenroos, S., Ahti, T., Lohtander, K., Myllys, L., Haikonen, V. (Eds.), 2011. Suomen jäkäläopas.
- 122 Kasvimuseo, Luonnontieteellinen keskusmuseo LUOMUS, Helsinki.
- 123 Stenroos, S., Velmala, S., Pyhälä, J., Ahti, T. (Eds.), 2015. Suomen rupijäkälät. Luonnontieteellinen
- 124 keskusmuseo LUOMUS, Helsinki.
- Stokland, J.N., Siitonen, J., Jonsson, B.G., 2012. Biodiversity in dead wood. Cambridge University Press,

126 Cambridge, UK.

Tuovila, H., 2013. Sticky business – diversity and evolution of Mycocaliciales (Ascomycota) on plant exudate. University of Helsinki.